

1 **Autopsy, thanatopraxy, cemeteries and crematoria as hotspots of toxic organic**
2 **contaminants in the funeral industry continuum: A hidden health risk or myth?**

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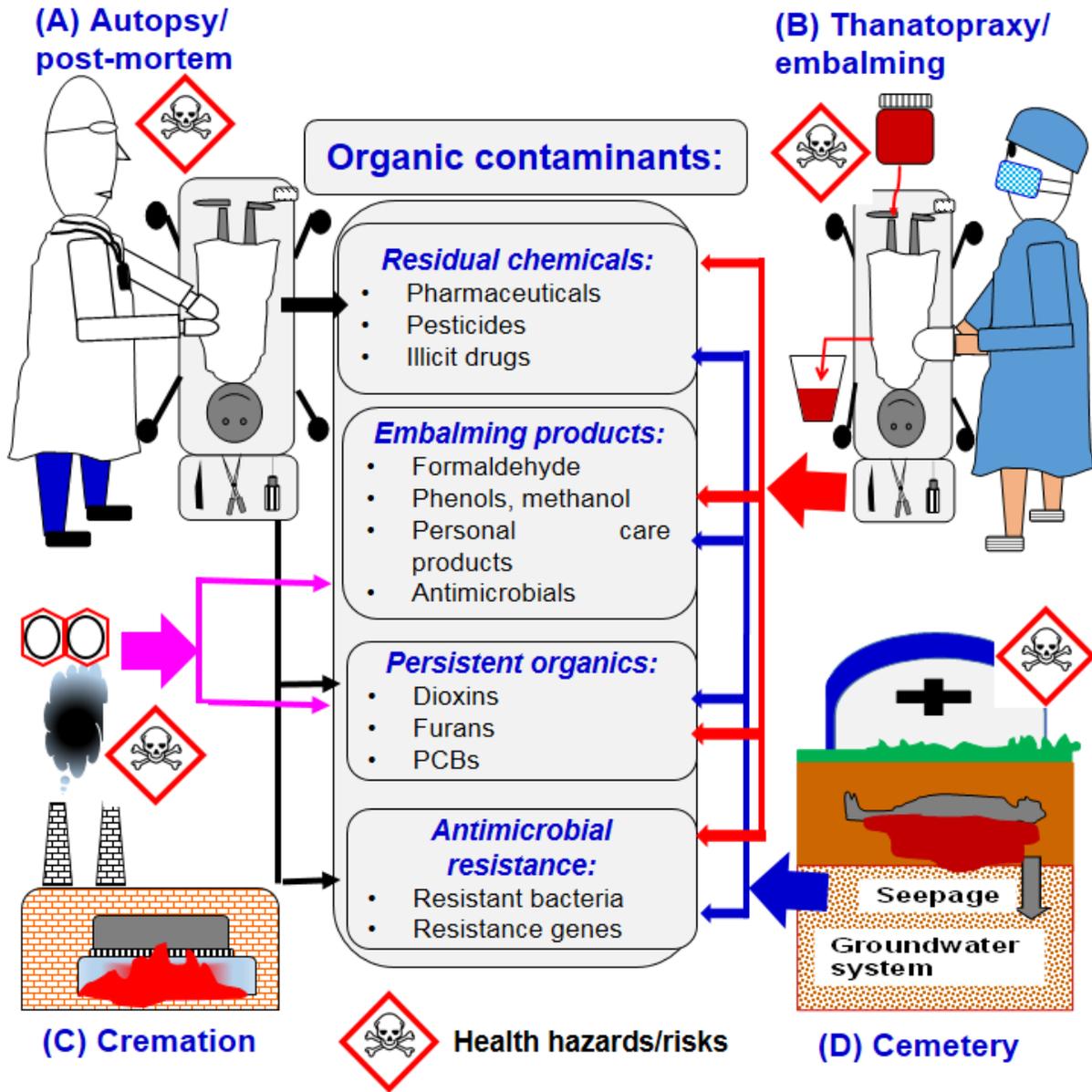
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13 ** Corresponding Author: wgwenzi@yahoo.co.uk; wgwenzi@agric.uz.ac.zw*

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18 **Statement of Novelty**

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21 The funeral industry is an under-studied and under-regulated potential hotspot of toxic
22 organic contaminants (TOCs). Yet comprehensive reviews tracking TOCs in the funeral industry
23 are missing. Embalming products, pharmaceuticals, personal care products, pesticides, persistent
24 organics and illicit drugs are detected. Autopsy, thanatopraxy, cemeteries and crematoria are
25 hotspots of TOCs. Health risks of TOCs, and flaws in current evidence and risk assessment
26 protocols are discussed. Risk assessment and mitigation, including regulatory, surveillance and
27 control systems are discussed. Knowledge gaps and emerging tools are highlighted. The findings
28 are comprehensive and significant, hence this review is a potential ‘hot’ paper on the topic.

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31 **Highlights**

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- 33 • The funeral industry is a hotspot source of toxic organic contaminants (TOCs).
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 - 35 • Yet comprehensive reviews of the occurrence and health risks of TOCs are scarce.
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 - 37 • Drugs, POPs, pesticides, antimicrobials, and embalming chemicals were detected.
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 - 39 • Human exposure, risk factors, health risks, and mitigation measures are discussed.
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 - 41 • Limitations of risk assessment protocols, and future directions are highlighted.
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60 **Abstract**

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The occurrence and health risks of toxic organic contaminants (TOCs) in the funeral industry are relatively under-studied compared to other industries. An increasing body of literature reports TOCs including emerging contaminants in the funeral industry, but comprehensive reviews of the evidence are still lacking. Hence, evidence was analysed to address the proposition that, the funeral industry constitutes several hotspot reservoirs of a wide spectrum of TOCs posing ecological and human health risks. Embalming products, persistent organic pollutants, synthetic pesticides, pharmaceuticals, personal care products and illicit drugs are detected. Human cadavers, solid wastes, wastewaters and air-borne particulates from autopsy, thanatopraxy care facilities (mortuaries, funeral homes), cemeteries and crematoria are hotspots of TOCs. Ingestion of contaminated water, and aquatic and marine foods constitute non-occupational exposure, while occupational exposure occurs via inhalation and dermal intake. Risk factors promoting exposure to TOCs include; unhygienic burial practices, poor solid waste and wastewater disposal, and weak and poorly enforced regulations. The health risks of TOCs are quite diverse, and include; (1) genotoxicity, endocrine disruption, teratogenicity and neurodevelopmental disorders, (2) development of antimicrobial resistance; (3) info-disruption via biomimicry, and (4) disruption of ecosystem functions and trophic interactions. Barring formaldehyde and inferential evidence, the epidemiological studies linking TOCs in the funeral industry to specific health outcomes are scarce. The reasons for the lack of evidence, and limitations of current health risk assessment protocols are discussed. A comprehensive framework for hazard identification, risk assessment and mitigation (HIRAM) in the funeral industry is proposed. The HIRAM includes regulatory, surveillance and control systems such as prevention and removal of TOCs. Future directions on the ecotoxicology of mixtures, behaviour, and health risks of TOCs are highlighted. The opportunities presented by emerging tools, including isotopic labelling, genomics, big data analytics (machine learning, artificial intelligence), and *in silico* techniques in toxicokinetic modelling are highlighted.

Keywords: Embalming products; emerging contaminants; human exposure pathways; mitigation framework; pharmaceuticals; thanatopraxy

91 **1 Introduction**

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The funeral industry continuum comprising of autopsy, thanatopraxy/embalming, funeral homes, and the subsequent burial of human cadavers in cemeteries, and cremation is a thriving global business. Similar to the health care system, the funeral industry is a putative hotspot reservoir of toxic organic contaminants (TOCs), posing human and ecological health risks. These TOCs include, embalming products from thanatopraxy, synthetic organic pesticides, persistent organic pollutants, and emerging contaminants originating from various sources. Despite being a seemingly obvious pollution hotspot, the occurrence and health risks of TOCs in the funeral industry are relatively under-studied compared to other industries. This is in contrast to substantial literature, including reviews, on the occurrence and health risks of TOCs including emerging contaminants in wastewater treatment systems, health care system, aquatic systems, pharmaceutical industries, and livestock production systems (Gwenzi and Chaukura, 2018; K'oreje et al., 2020; Michael et al., 2020). An earlier review discussed the several reasons accounting for this trend (Gwenzi, 2020). This include the fact that, the funeral industry is

106 classified under commercial service providers, which are less regulated and monitored compared
107 to other industrial sectors such as the health care system (Davidson and Benjamin, 2006).

108 An increasing body of literature exists on inorganic, organic and microbial pollution from
109 the funeral industry, particularly cemeteries, and to a less extent, embalming and cremation.
110 Specifically, a number of studies have investigated the following: (1) inorganic contaminants such
111 as metals and nutrients, and organic matter (i.e., COD, BOD) in surface and drainage water from
112 cemeteries (Spongberg and Becks, 2000; Fielder et al., 2012, Fineza et al., 2014; Vaezihir and
113 Mohammadi, 2016), (2) microbial contaminants, including bacteria and viruses, and human
114 pathogens in thanatopraxy and cemeteries (Davidson and Benjamin, 2006; Carstens et al., 2014;
115 Abia et al., 2018). Other studies investigated the occurrence and health risks of formaldehyde in
116 thanatopraxy and cemeteries (Aronson et al., 2004; Zume et al., 2011; Varlet et al., 2019), while
117 others detected persistent organic pollutants such as dioxins and furans in crematoria (Takeda et
118 al., 2000, 2004; Wang et al., 2003). More recently, an increasing number of studies have detected
119 several emerging contaminants, including pharmaceuticals, illicit drugs and personal care products
120 in funeral homes and cemeteries (Paíga and Delerue-Matos, 2016; Fielder et al., 2017; McBean et
121 al., 2018; Kleywegt et al., 2019).

122 Despite the increasing evidence on the occurrence of a wide spectrum of TOCs in the
123 funeral industry, comprehensive reviews on the subject are still lacking. The few earlier reviews
124 were limited to toxic metals, nutrients, pathogenic and indicator microorganisms (bacteria, viruses)
125 and organic matter in cemeteries (Demiryurek et al., 2002; Oliveira et al., 2013; Żychowski, 2012;
126 Żychowski and Bryndal, 2015. Barring cemeteries, earlier reviews excluded TOCs in other
127 compartments in the funeral industry, specifically, autopsy, thanatopraxy, funeral homes and
128 cremation. In addition, earlier reviews excluded several emerging contaminants recently reported
129 in literature (e.g., Paíga and Delerue-Matos, 2016; Fielder et al., 2017; McBean et al., 2018). This
130 is because until now, such evidence was not available to enable comprehensive reviews on the
131 subject. Therefore, the current review traces the occurrence, behaviour, exposure pathways, risk
132 factors, and health risks of TOCs in the funeral industry (Figure 1).

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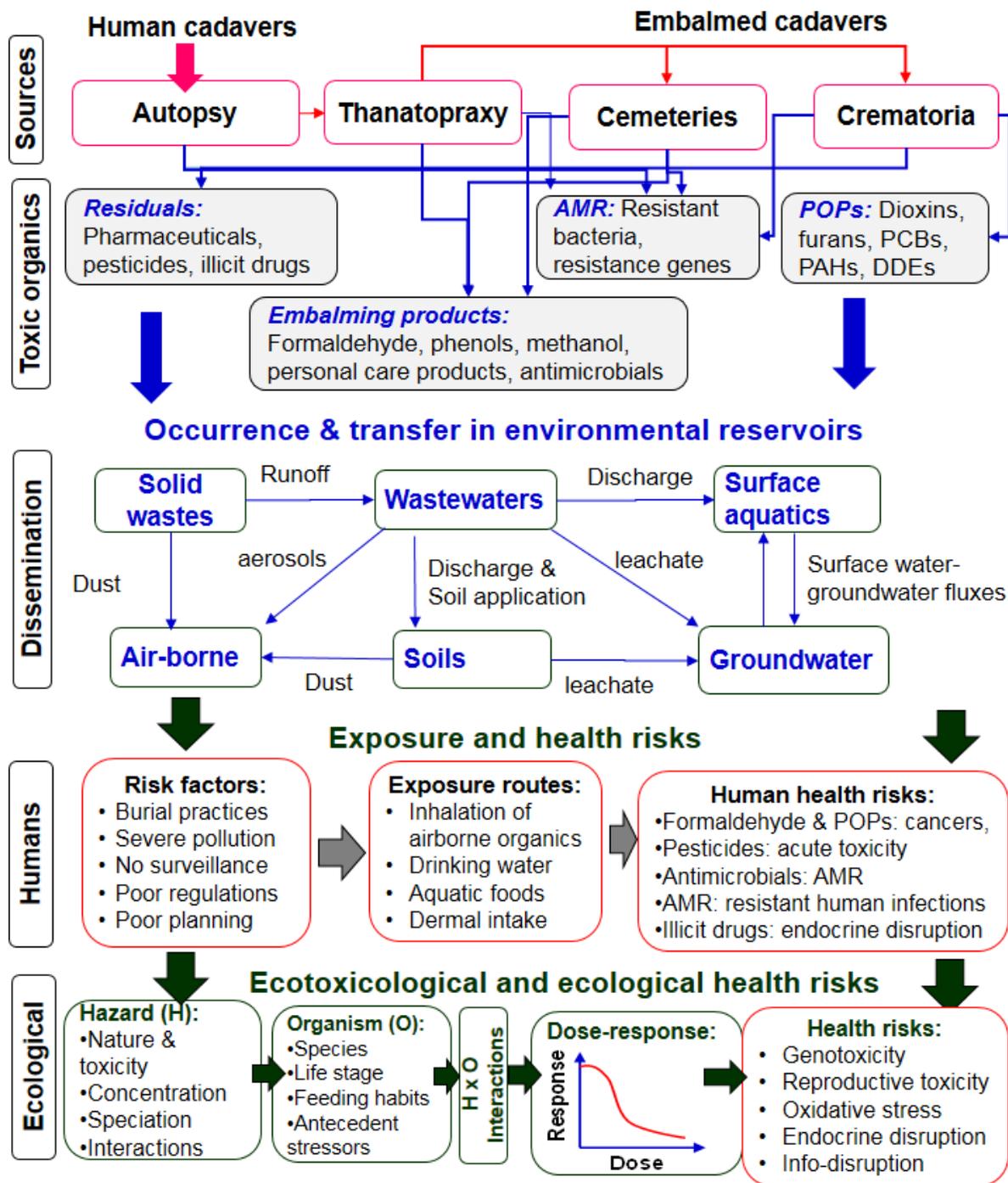


Figure 1. Summary depiction of the nature, occurrence, dissemination and health risks of organic contaminants in the funeral industry and cemeteries.

The review posits that, a wide spectrum of TOCs occur in the funeral industry, and their occurrence, coupled with the various exposure pathways and risk factors, poses significant ecological and human health risks. In summary, compared to the few earlier reviews focusing on

142 inorganic and microbial contamination in single compartments specifically cemeteries, the current
143 review makes a number of contributions. First, it considers funeral industry as a continuum
144 consisting of autopsy, thanatopraxy in funeral homes, cemeteries and crematoria. Second, it
145 addresses a broad spectrum of TOCs, including embalming products, synthetic organic pesticides,
146 persistent organic pollutants, and emerging contaminants, specifically pharmaceuticals, personal
147 care products and illicit drugs. Exposure pathways and risk factors predisposing human health to
148 TOCs in developing regions particularly Africa are discussed. The weak evidence on health risks,
149 and limitations of existing health risk assessment protocols are highlighted. A comprehensive
150 framework for hazard identification, risk assessment and mitigation (HIRAM), including
151 regulatory, policy, surveillance and control systems is proposed. Finally, future research directions,
152 including knowledge gaps and the need to harness emerging research tools are highlighted.

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154 **2 Materials and methods**

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156 Literature was retrieved from scholarly databases, including; Clarivate's Web of Science,
157 Scopus®, ResearchGate®, Google Scholar®, ScienceDirect®) and university thesis repositories,
158 among others. The searches were conducted using Boolean search techniques as described in
159 earlier studies (Gwenzi et al., 2017; Gwenzi and Chaukura, 2018). Briefly, to retrieve articles with
160 both the funeral industry and TOCs, the Boolean search technique based on the use of 'AND'/'OR'
161 was used. The search terms used were: 'funeral industry', 'funeral service', 'funeral sector',
162 'funeral service provider', 'funeral homes', 'embalming', 'morgues' 'funeral parlours',
163 'thanatopraxy', 'mortuaries', 'autopsy', and 'necropsy'. The other terms used were; 'human
164 cadaver', 'human corpse', 'human dead bodies', 'human remains', 'graves', 'gravesites',
165 'graveyards', 'burial sites', 'tombs', 'cemeteries', 'cremation' and 'crematorium(s)/crematoria'.

166 The search terms used for TOCs were: 'organic contaminants', 'emerging contaminants',
167 'contaminants of emerging concern', 'synthetic organic contaminants', and names of specific
168 groups of TOCs, including; 'embalming fluids/chemicals/products', 'pesticides', 'persistent
169 organic pollutants', 'pharmaceuticals', 'antimicrobials', 'antibiotics', 'anti-retrovirals',
170 'antifungals', 'disinfectants', 'personal care products', 'drugs', 'therapeutic drugs', 'fragrances',
171 'perfumes', 'anti-odorants', anticoagulants, and 'penetrating agents'. Literature on TOCs from
172 forensic studies conducted during suicide or homicide post-mortems was also included. The
173 reference lists of relevant articles were also manually searched for additional articles. The retrieved
174 articles were examined, and the key findings on the occurrence and health risks of TOCs were
175 summarized and tabulated (Tables 1 and 2). In the current study, a quantitative analysis based on
176 bibliometric and meta-analytical approaches was not feasible because such analyses require a large
177 dataset which was not available. Thus, given the limited dataset, the current review was limited to
178 qualitative analysis.

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180 **3 Occurrence of TOCs in the funeral industry**

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182 **3.1 The funeral industry as a continuum of TOCs**

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184 In the current review, the funeral industry is conceptualized as a continuum of various
185 hotspot compartments of TOCs including autopsy, thanatopraxy/funeral homes, and ultimately
186 cemeteries, and crematoria (Figure 1). On one hand, TOCs such as embalming products,
187 antimicrobials and disinfectants are added at various points along the continuum. On the other

188 hand, TOCs are released from various reservoirs including human cadavers, solid wastes, body
189 fluids, wastewaters, and air-borne particulates and aerosols. Six broad groups of TOCs were
190 detected: (1) embalming products, (2) pharmaceuticals, (3) personal care products, (4) persistent
191 organic contaminants, (5) synthetic organic pesticides, and (6) illicit drugs (Figure 1, Table 1).
192 Human cadavers, solid wastes, wastewaters and air-borne particulates from autopsy, thanatopraxy
193 care facilities, cemeteries and crematoria are the hotspots of TOCs (Figure 1).

194 Pharmaceuticals and their metabolites occur as residuals from their widespread
195 applications in health care systems and as over-the-counter drugs (Kleywegt et al., 2019).
196 Embalming products, personal care products, disinfectants and antimicrobials originate from
197 autopsy and thanatopraxy care facilities, including mortuaries and funeral homes (Varlet et al.,
198 2019). Synthetic organic pesticides and illicit drugs originate from human cadavers following
199 accidental and deliberate exposures in suicide and homicide cases (Pilgrim et al., 2011). Persistent
200 organic pollutants (POPs) such as dioxins, furans and polychlorinated biphenyls (PCBs) are
201 released during cremation (Takeda et al., 2000, 2004; Mari and Domingo, 2016). Persistent
202 pesticides and their metabolites (e.g., Lindane, DDE) bioaccumulate in fatty tissue (ATSDR,
203 2019). Hence, they are also released from human bodies via body fluids (e.g., urine), and faecal
204 matter, following long-term exposure (ATSDR, 2019; Kleywegt et al., 2019).

205 Besides human cadavers, the funeral industry generates solid wastes, wastewaters and
206 airborne particulates and aerosols (Gwenzi, 2020). For example, autopsy, forensic studies and
207 thanatopraxy involve the manipulation and dissection of human cadavers, including sampling of
208 human tissue and body fluids (Nwanyanwu et al., 1989; Davidson and Benjamin, 2006).
209 Aspiration from swollen organs may lead to sudden release of body fluids (Davidson and
210 Benjamin, 2006). Excessive gurgling and frothing through the mouth and noses, and leakage of
211 faecal matter via the anus and dissected gut system may also occur (Creely, 2004; Davidson and
212 Benjamin, 2006).

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Table 1. Summary of conventional and emerging organic contaminants reported in the autopsy, thanatopraxy, cemeteries and crematoria

Organic contaminants	Nature, concentration and reference	Remarks
A: Embalming products:		
(1) Formaldehyde/formalin	Data from Young et al. (2002) show that: (1) 9 mg/L of formaldehyde reported in water from a cemetery in London, UK. (2) 0.1 L of formaldehyde is leached per human cadaver embalmed. 40 mg/L is expected in leachate in first year, dropping to about 5 mg/L after 10 years.	Formaldehyde originates from embalming fluids and also medium density fibreboard used to make coffins.
	87100 ng/L (median) and 561000 ng/L (maximum) of formaldehyde were reported in embalming wastewater in Ontario, Canada (Kleywegt et al., 2019)	Samples were collected during active embalming.
(2) Triclosan	18000 ng/L (median) and 505000 ng/L (maximum) detected in embalming wastewater in Ontario, Canada (Kleywegt et al., 2019)	Triclosan is used as an antimicrobial in household disinfectants
(3) Methanol, glycol, phenol and glutaraldehyde	These are common chemicals used in varying percentages in various embalming products ((NFDA, 1995)	No data is available on their concentrations in environmental media
B: Persistent organic pollutants (dioxins, furans and PCBs):		
(1) Polychlorinated dibenzodioxins (PCDDs) and polychlorinated dibenzofurans (PCDFs)	Data in crematoria in Japan showed the following results (Takeda et al., 2000, 2014): (1) PCDDs/PCDFs concentrations of 2.2-290 ng/N m ³ , giving total equivalents (TEQ) of 0.0099-6.5 ng TEQ/N m ³ . (2) Estimated total PCDDs/PCDFs emission was 8.9 gTEQ/year	Concentration data were measured to 10 crematoria, while total emissions were estimated based on all crematoria in Japan. Emissions were highest in the first 20 minutes of cremation.
(2) Polychlorinated biphenyls (PCBs)	16 ng/L (mean) and 290 ng/L (maximum) total PCBs detected in embalming wastewater (Kleywegt et al., 2019)	Samples were collected during active embalming.

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Table 1. (contd.).

Organic contaminants	Nature, concentration and reference	Remarks												
C: Synthetic organic pesticides:														
(1) Methyl parathion	24 ug/mL of methyl parathion detected in post-mortem blood of a 21 year old woman (Tsoukali et al., 2004).	Person died from fatal poisoning by suicide using intravenous injection of methyl-parathion												
(2) Imidacloprid [1-(6-chloro-3pyridylmethyl)-N-nitroimidazolidin-2-ylideneamine]	Concentrations (ug/ml) of Imidacloprid detected in post-mortem blood of two males were (Proença et al., 2005): 33-year old: 12.5 (blood), 13.6 (kidney), 9.9 (liver), 20.6 (lung) 66-year old: 2.05 (blood), 2.5 (kidney), 1.01 (kidney), 8.8 (lung)	Imidacloprid was detected using liquid chromatography/mass spectrometry (LC/MS) in post-mortem blood samples from suicide victims in Portugal.												
(3) Methomyl (S-methyl-N-[(methylcarbamoyl)oxy]thioacetimidate)	Ranges and (mean) concentration (mg/L) of methomyl, a carbamate pesticide were: Blood: 5.6-63.5 (27.4), vitreous humour: 1.4-7.0, liver 0.1-1.2 (0.6), kidney: 0.2-2.8 (1.2), brain 0.07-0.31 (0.17) (Tsatsakis and Tsakalof, 1996)	Methomyl concentrations were determined in autopsy human material from 8 cases of fatal human poisoning in Crete, Greece.												
(4) Lindane	83 ng/L detected in embalming wastewater (Kleywegt et al., 2019)	Lindane is a persistent pesticide long-banned in several countries												
(5) pp-DDE (1,1-dichloro-2,2-bis(p-chlorophenyl)ethylene)	11.0 ng/L (mean) and 2300 ng/L (maximum) detected in embalming wastewater (Kleywegt et al., 2019).	DDE is a metabolite of the degradation of DDT, a persistent pesticide long-banned in several countries												
D: Personal care products:														
Oxybenzone	227.2 ng/L (median) and 8,160 ng/L (maximum) were detected in embalming wastewater (Kleywegt et al., 2019).	Oxybenzone is commonly used in sunscreens												
DEET (N,N-Diethyl-m-toluamide)	53.6 ng/L (mean) and 3,280 ng/L (maximum) were detected in embalming wastewater (Kleywegt et al., 2019).	DEET is a common ingredient in insect repellants.												
Nonyl phenol and its metabolites	Median and maximum concentrations (ng/L) of nonyl phenols in embalming wastewaters were (Kleywegt et al., 2019): <table border="1" style="margin-left: auto; margin-right: auto;"> <thead> <tr> <th></th> <th>Med.</th> <th>Max.</th> </tr> </thead> <tbody> <tr> <td>4-Nonylphenol</td> <td>450</td> <td>14000</td> </tr> <tr> <td>4-Nonylphenol Monoethoxylate</td> <td>230</td> <td>7200</td> </tr> <tr> <td>4-Nonylphenol Diethoxylate</td> <td>210</td> <td>5500</td> </tr> </tbody> </table>		Med.	Max.	4-Nonylphenol	450	14000	4-Nonylphenol Monoethoxylate	230	7200	4-Nonylphenol Diethoxylate	210	5500	Nonyl phenol is used as a surfactant, while Nonylphenol monoethoxylate and -Nonylphenol diethoxylate are its metabolites
	Med.	Max.												
4-Nonylphenol	450	14000												
4-Nonylphenol Monoethoxylate	230	7200												
4-Nonylphenol Diethoxylate	210	5500												

229 **Table 1.** (contd.)
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Organic contaminants	Nature, concentration and reference	Remarks																											
<i>E: Pharmaceuticals and illicit drugs:</i>																													
(1) Several therapeutic drugs	<p>Concentration range (mean) of drugs in surface and drainage water samples (Fiedler et al., 2017):</p> <table border="0"> <tr> <td></td> <td style="text-align: center;"><i>Surface water:</i></td> <td style="text-align: center;"><i>Drainage:</i></td> </tr> <tr> <td>Metoprolol</td> <td style="text-align: center;">2230 ng/L</td> <td style="text-align: center;">23 ng/L</td> </tr> <tr> <td>Carbamazepine</td> <td style="text-align: center;">43–418 (222)</td> <td style="text-align: center;">10–225 (85)</td> </tr> <tr> <td>Hydrochlorothiazide</td> <td style="text-align: center;">Detected</td> <td style="text-align: center;">Detected</td> </tr> <tr> <td>Diclofenac</td> <td style="text-align: center;">129–574 (314)</td> <td style="text-align: center;">Not detected</td> </tr> <tr> <td>Atenolol</td> <td style="text-align: center;">57–301</td> <td style="text-align: center;">Not detected</td> </tr> <tr> <td>Naproxen</td> <td style="text-align: center;">41-81</td> <td style="text-align: center;">Not detected</td> </tr> <tr> <td>Ibuprofen</td> <td style="text-align: center;">Not detected</td> <td style="text-align: center;">Detected</td> </tr> </table>		<i>Surface water:</i>	<i>Drainage:</i>	Metoprolol	2230 ng/L	23 ng/L	Carbamazepine	43–418 (222)	10–225 (85)	Hydrochlorothiazide	Detected	Detected	Diclofenac	129–574 (314)	Not detected	Atenolol	57–301	Not detected	Naproxen	41-81	Not detected	Ibuprofen	Not detected	Detected	<p>Focused on cemeteries in regions with relatively high consumption of medicines in Germany. 12 pharmaceuticals tested were detected in surface water compared to just 5 in drainage. Concentrations were generally higher in surface than drainage water samples, indicating attenuation/removal in the soil.</p>			
	<i>Surface water:</i>	<i>Drainage:</i>																											
Metoprolol	2230 ng/L	23 ng/L																											
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Naproxen	41-81	Not detected																											
Ibuprofen	Not detected	Detected																											
(2) Several active pharmaceuticals	<p>Median (med.) and maximum (max.) concentrations (ng/L) were (Kleywegt et al., 2019):</p> <table border="0"> <tr> <td></td> <td style="text-align: center;">Med.</td> <td style="text-align: center;">Max.</td> </tr> <tr> <td>Caffeine</td> <td style="text-align: center;">76650</td> <td style="text-align: center;">14,200,000</td> </tr> <tr> <td>Ketoprofen</td> <td style="text-align: center;">56</td> <td style="text-align: center;">900</td> </tr> <tr> <td>Acetaminophen (paracetamol)</td> <td style="text-align: center;">15,520</td> <td style="text-align: center;">1,720,000</td> </tr> <tr> <td>Hydrocortisone</td> <td style="text-align: center;">1,524</td> <td style="text-align: center;">22,000</td> </tr> <tr> <td>Lidocaine</td> <td style="text-align: center;">39.68</td> <td style="text-align: center;">21,000</td> </tr> <tr> <td>Naproxen</td> <td style="text-align: center;">389</td> <td style="text-align: center;">310,000</td> </tr> <tr> <td>Ibuprofen</td> <td style="text-align: center;">286</td> <td style="text-align: center;">980,000</td> </tr> <tr> <td>Ciprofloxacin</td> <td style="text-align: center;">43.6</td> <td style="text-align: center;">93,000</td> </tr> </table>		Med.	Max.	Caffeine	76650	14,200,000	Ketoprofen	56	900	Acetaminophen (paracetamol)	15,520	1,720,000	Hydrocortisone	1,524	22,000	Lidocaine	39.68	21,000	Naproxen	389	310,000	Ibuprofen	286	980,000	Ciprofloxacin	43.6	93,000	<p>Wastewater samples collected during active embalming in Ontario, Canada. This is probably the first study to determine emerging contaminants in embalming wastewater. The concentrations are generally higher than those reported for other environmental media in literature.</p>
	Med.	Max.																											
Caffeine	76650	14,200,000																											
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Ciprofloxacin	43.6	93,000																											
(3) Several antibiotics	<p>Tetracyclines, Clavulanic acid, Vancomycin, Novobiocin, Butirosine and Neomycin, Streptomycin; Penicillins, Cephalosporins, Ansamycin and 12-, 14- and 16-membered Macrolides (Abia et al., 2019).</p>	<p>Detected on soils from cemeteries in South Africa. Biosynthesis was also observed in the study.</p>																											
(4) Several drugs	<p>The drugs amitriptyline (1811 pg/g in hair, 29.8 ug/g in liver), nortriptyline (43 pg/g in hair, 3.6 ug/g in liver) and bromazepam (740 pg/g in hair) were detected (Gaillard et al., 2011).</p>	<p>Autopsy was conducted 8 months after death of a one-month old girl due to homicidal poisoning.</p>																											
(5) MDMA “ecstasy” (3,4-methylenedioxy-methamphetamine)	<p>The concentrations of MDMA detected in post-mortems in Australia were (Pilgrim et al., 2011): 0.02 to 3.5 mg/L. Study based on national database of the National Coroners Information System, Australia.</p>	<p>Other illicit (amphetamines, codeine, cocaine, morphine) and therapeutic drugs (paracetamol, moclobemide) were detected in the order of mg/L.</p>																											

232 **Table 1.** (contd.)
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Organic contaminants	Nature, concentration and reference	Remarks
(5) Several illicit and pharmaceutical drugs	Data from 24, 876 forensic autopsies in Sweden detected over 25 drugs and their metabolites in blood (Jones and Homgren 2009). Mean concentrations (mg/L) of the top 10 most common drugs were: ethanol (1430), Acetaminophen/paracetamol (25.5), Citalopram (0.72); Diazepam (0.23), Zopiclone (0.3), Morphine (0.3), codeine (0.32), Tramadol (2.64); Alimemazine (0.54); Dextropropoxyphene (2.0)	Several other illicit drugs and therapeutic drugs were also detected in orders of mg/L. Data are based on analysis done on femoral venous blood samples for forensic purposes.
(6) Cocaine	In 49 autopsy cases covering a period of 11 years cocaine concentration range from 0.01-3.0 mg/L with a median of 0.1 mg/L (Pilgrim et al., 2013). Cocaine metabolites w also detected.	Based data from the National Coroners Information System database for all deaths in Victoria, Australia for 2000 to 2011.

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235 Subsequent washing, dressing and bagging of cadavers generate solid wastes and
236 wastewaters. Data show that about approximately 13.2 litres of embalming fluid is required to
237 embalm an average adult (Chiapelli and Chiapelli, 2008). Thus, for a country with about 2 million
238 people embalmed each year (e.g., USA), approximately 26.5 million litres of embalming fluid is
239 required. This generate an equivalent or even more amounts of body fluids and wastewaters that
240 are directly released into the on-site sanitation systems, municipal wastewater systems or the
241 environment. TOCs in solid wastes often end up in either sanitary landfills in developed counties
242 or non-sanitary waste dumps in developing countries, where they undergo further dissemination
243 (Figure 2). Human cadavers are ultimately buried in cemeteries or cremated. TOCs in human
244 cadavers can be released into the environment via cadaver slurries and seepages during
245 decomposition (Abia et al., 2018; 2019). Cremation of cadavers releases TOCs as volatiles or in
246 air-borne particulates (Da Cruz et al., 2012). Taken together, the funeral industry is a continuum
247 of several hotspot sources of a wide range of TOCs, including emerging contaminants. Yet despite
248 being a hotspot of TOCs, the funeral industry is relatively under-studied relative to comparative
249 industries such as health care systems (Horton, 2003; Davidson and Benjamin, 2006; Miranda,
250 2016; Kuchniski et al., 2019). The reasons accounting for this trend are discussed in detail
251 elsewhere (Gwenzi, 2020), and they include: (1) regulatory frameworks that classify funeral
252 industry under commercial service providers, thereby implying low health risk, (2) socio-cultural
253 factors, (3) risk of contracting human infections, (4) Matthew or bandwagon effect, and (5) limited
254 interest by both researchers and funding organization caused by negative perceptions and attitudes.

255

256 **3.2 Nature and sources of TOCs**

257

258 **3.2.1 Pharmaceuticals**

259

260 Pharmaceuticals have been reported in soils (Abia et al., 2019), surface water and
261 groundwater from cemeteries (Fiedler et al., 2011; Paíga and Delerue-Matos, 2016), human
262 cadavers (Gaillard et al., 2011), and embalming wastewaters (Kleywegt, 2019) (Table 1). Abia et
263 al. (2019) detected over ten pharmaceuticals in cemetery soils in South Africa, including;
264 Ansamysin, Cephalosporins, Penicillins, Streptomycin, Butirosine, Neomycin, Novobiocin,
265 Vancomycin, Clavulanic acid, Tetracyclines and 12-, 14- and 16-membered Macrolides. In an
266 autopsy of a one-month old girl who died due to homicidal poisoning, the following drugs were
267 detected: (1) Amitriptyline (1811 pg/g in hair, 29.8 ug/g in liver), (2) Nortriptyline (43 pg/g in hair,
268 3.6 ug/g in liver), and (3) Bromazepam (740 pg/g in hair) (Gaillard et al., 2011).

269 In cemeteries in Germany, Fielder et al. (2017) showed that five samples contained
270 Hydrochlorothiazide, and one each had Metoprolol (23 ng/L) and traces of Ibuprofen (Fielder et
271 al., 2017). In the same study, eight out of the 12 pharmaceuticals tested were detected in surface
272 water samples compared to just five in drainage water (Table 1). The concentrations were higher
273 in surface than drainage water, as evidenced by Metoprolol which had a concentration of 2230
274 ng/L. The lower concentrations in drainage water relative to surface water point to attenuation or
275 degradation mechanisms in the soil. Carbamazepine was detected in the highest concentration
276 (mean: 225 ng/L) in seven out of the total 12 drainage water samples analysed. Carbamazepine is
277 chemically stable, highly persistent and has a low octanol/water partitioning coefficient ($\log K_{ow}$)
278 of about 2.25 to 2.45 (Jones et al. 2005; Fiedler et al., 2017). Hence, it undergoes limited adsorption
279 onto solid matrix such as soils and sediments in water/wastewater treatment systems and aquatic
280 systems (Löffler et al. 2005; Fatta-Kassinos et al., 2011; Stuart et al., 2012; Fiedler et al., 2017).

281 For example, concentrations of carbamazepine as high as 3600 ng/L have been reported in
282 groundwater (Stuart et al., 2012), while 42 ng/L was detected in drinking water (Vulliet and Cren-
283 Olive 2011). Due to its persistence and limited adsorption, carbamazepine is often applied as a
284 tracer for anthropogenic contamination in aquatic systems (Clara et al., 2004; Ruzicka et al., 2011).

285 In Portugal, nonsteroidal anti-inflammatory analgesics (e.g., Salicylic Acid), antibiotics
286 and psychiatric drugs (e.g., Carbamazepine, Fluoxetine, Nimesulide) were investigated in two
287 groundwater samples each from a total of five cemeteries (Paíga and Delerue-Matos, 2016).
288 Salicylic acid, Carbamazepine, Ketoprofen and Ibuprofen were detected in all ten samples. In cases
289 where they were detected, Ibuprofen, Acetaminophen (Paracetamol), Ketoprofen and Sertraline
290 their concentrations were below the method detection limits for the UHPLC–MS/MS triple-
291 quadrupole mass spectrometer (Paíga and Delerue-Matos, 2016). Salicylic acid and carbamazepine
292 were detected in all 10 samples, fluoxetine in 8 samples, while Nimesulide was detected in one
293 sample. The mean (\pm RSD) concentrations (ng/L) for the four pharmaceuticals decreased in the
294 order: (1) Salicylic Acid: $37.7 \pm 4.85 - 71.0 \pm 1.48$, (2) Carbamazepine: $20.0 \pm 0.58 - 23.8 \pm 1.91$,
295 (3) Nimesulide: 9.24 ± 9.60 , and (4) Fluoxetine: $1.90 \pm 1.24 - 1.97 \pm 4.79$ (Paíga and Delerue-Matos,
296 2016). Salicylic Acid showed more variation than carbamazepine and fluoxetine.

297 In Ontario, Canada, high concentrations of pharmaceuticals were detected in embalming
298 wastewater (Kleywegt et al., 2011; Table 1). The concentrations (median; maximum) (ng/L) were:
299 (1) Ketoprofen (56, 900), (2) Acetaminophen (Paracetamol) (15,520; 1,720,000), (3)
300 Hydrocortisone (1,524; 22,000), (4) Lidocaine (39.68; 21,000), (5) Naproxen (389; 310,000), (6)
301 Ibuprofen (286; 980,000) and (7) Ciprofloxacin (43.6; 93,000). To the author's knowledge, this is
302 the first study to investigate pharmaceuticals in embalming wastewater, and the concentrations
303 reported are amongst the highest values reported in literature so far. Pharmaceuticals are generally
304 not persistent in the environment, but are regarded as 'pseudo-persistent' (Ebele et al., 2017). This
305 is because of their continual use and excretion, coupled with their low removal in water/wastewater
306 treatment facilities (Ebele et al., 2017).

307

308 **3.2.2 Embalming products**

309

310 Embalming products are complex mixtures of cosmetic products, penetrating agents, anti-
311 coagulants, cleaners, and antimicrobials such as methanol, formaldehyde, glycol, phenol and
312 glutaraldehyde (Kleywegt et al., 2019; Varlet et al., 2019). Embalming products vary considerably
313 in terms of brands and composition with respect to per cent formaldehyde (1-40%), propylene
314 glycol (1-100%) phenol (2-40%), glutaraldehyde (10-30%), methanol (1-100%) and Triclosan
315 (Kleywegt et al., 2018). Formaldehyde, glutaraldehyde and phenol have been detected in
316 occupational settings and in the environment (Aronson et al., 2004; Zume et al., 2011; Kleywegt
317 et al., 2017). For example, 87100 ng/L (median) and 561000 ng/L (maximum) of formaldehyde
318 were detected in embalming wastewater in Ontario, Canada (Kleywegt et al., 2019). In Nigeria,
319 Tume (2011) showed that 57% of groundwater samples close to graves had phenol concentrations
320 (0.1-2.6 mg/L) which were more than 2 orders of magnitude higher than the Nigerian Standard for
321 Drinking Water Quality of 0.001 mg/L.

322 Due to the toxicity of formaldehyde to both the environment and practitioners, a wide range
323 of synthetic embalming chemicals have been developed and are currently used for thanatopraxy
324 (Waschke et al. 2019; Varlet et al., 2019). A recent review by Varlet et al. (2019) presents over 30
325 organic formulations for embalming fluids, which are used as alternatives to formaldehyde.
326 Interestingly, some antimicrobials, pharmaceuticals, personal care products and anti-coagulants

327 are emerging contaminants (Eljarrat et al., 2012; Gwenzi and Chaukura, 2019). Yet besides the few
328 antimicrobials detected so far, data on penetrating agents, anticoagulants, moisturizers and
329 fragrances and dyes used in the new embalming products are lacking (Varlet et al., 2019).

330

331 **3.2.3 Personal care products**

332

333 Personal care products include antimicrobials (e.g., triclosan), insect repellents (e.g.,
334 N,Ndiethyl-m-toluamide also known as DEET), sunscreen agents such as ultraviolet filters (e.g.,
335 oxybenzone) and surfactants (e.g., nonyl phenols), among others (Gros et al., 2008; Kleywegt et
336 al., 2019). Triclosan, DEET, oxybenzone, and nonyl phenols and their metabolites have been
337 detected in wastewater from embalming process (Kleywegt et al., 2017; Table 1). For example,
338 18000 ng/L (median) and 505000 ng/L (maximum) of triclosan used in both embalming products
339 and personal care products were detected. In the same study, 53.6 ng/L (mean) and 3,280 ng/L
340 (maximum) of DEET were detected (Kleywegt et al., 2019). Nonyl phenols, and their metabolites
341 were also detected with the following concentrations (mean; maximum): Nonylphenol (450; 14000
342 ng/L), 4-Nonylphenol Monoethoxylate (230; 7200 ng/L), and 4-Nonylphenol Diethoxylate (210;
343 5500 ng/L). Personal care products are ubiquitous due to their widespread use in household
344 products including; shampoos, lotions, soaps deodorants, detergent, toothpastes and disinfectants
345 and toothpastes (Peck, 2006; Gros et al., 2008). Personal care products are complex mixture of
346 active, non-biodegradable and pseudo-persistent emerging contaminants (Rasheed et al., 2019).
347 Due to their lipophilicity, personal care products tend to bioaccumulate in the fatty tissue (Ebele
348 et al., 2017). In the literature, a diverse range of personal care products is reported including
349 parabens and solvents (Sorensen et al., 2015; Gwenzi and Chaukura, 2018). Moreover, given rapid
350 developments in the field of synthetic chemistry, one may also expect that several personal care
351 products are yet to be documented in literature. Thus, the few groups of personal care products
352 highlighted here only represents a small fraction of a large group of such emerging contaminants.

353

354 **3.2.4 Synthetic organic pesticides**

355

356 Pesticides and their metabolites have been detected in embalming wastewaters (Klewegt
357 et al., 2019), and autopsy and forensic studies (Tsoukali et al., 2005; Proença et al., 2005). Lindane,
358 a persistent pesticide was detected with a concentration of 83 ng/L) in embalming wastewater
359 (Kleywegt et al., 2019). 1,1-dichloro-2,2-bis(p-chlorophenyl) ethylene (pp-DDE) was detected in
360 embalming wastewaters with mean and maximum concentrations of 11.0 ng/L and 2300 ng/L,
361 respectively (Kleywegt et al., 2019). DDE is a metabolite of DDT (1,1'-(2,2,2-Trichloroethane-
362 1,1-diyl)bis(4-chlorobenzene), a highly persistent pesticide long-banned in several countries
363 (ATSDR, 2019; Kleywegt et al., 2019). The occurrence of DDE in Canada was highly unexpected
364 because its use was banned nearly 50 years ago in 1972 (Kleywegt et al., 2019). DDE was
365 attributed to high concentrations of DDT in immigrants from developing countries where DDT is
366 still in use to some extent (Kleywegt et al., 2019). Given the persistence and bioaccumulation of
367 DDT and its metabolites (ATSDR, 2019), a possibility also exists that some of the cadavers
368 embalmed at the time of sampling could have been exposed to DDT before its ban in 1972.

369 Post-mortem and forensic studies have also reported the following pesticides: (1) Methyl
370 parathion (Tsoukalis et al., 2004), (2) 1-(6-chloro-3pyridylmethyl)-N-nitroimidazolidin-2-
371 ylidenamine, commonly known as Imidacloprid (Proença et al., 2005), and (3) S-methyl-N-
372 [(methylcarbamoyl)oxy]thioacetimidate also known as Methomyl (Tsatsakis and Tsakalof, 1996,

373 Table 1). These three cases involved deliberate fatal poisoning (suicide) through intravenous
374 injection or ingestion. Thus, high concentrations of pesticides in human cadavers and their body
375 fluids, solid wastes and wastewaters may occur in cases of suicide, homicide and accidental intake
376 (Gunnell and Eddleston, 2003; Jones and Homgren, 2009; Pilgrim et al., 2011; Freire and Koifman,
377 2012). In developing regions including Africa, human deaths and suicides due to fatal exposure to
378 pesticides are prevalent in the agricultural sector (Gunnell and Eddleston, 2003). Yet limited data
379 exist documenting the occurrence and health risks of pesticides in the funeral industry in Africa.
380

381 **3.2.5 Persistent organic pollutants**

382
383 Polychlorinated dibenzodioxins (PCDDs), polychlorinated dibenzofurans (PCDFs) and
384 polychlorinated biphenyls (PCBs) are persistent organic pollutants (POPs) formed during the
385 combustion of chlorinated materials such as plastics in prosthetics, body bags, caskets and fuels
386 such as wood containing chlorine (Kuchnicki, 2019). PCDDs, PCDFs and PCBs have been
387 detected in crematoria in several countries, including; China, United Kingdom, the USA and Japan
388 (Wang et al., 2003; Takeda et al., 2000, 2014; Dummer et al., 2003; Guttman et al., 2012; Xue et
389 al., 2016. Table 1). A study based on ten crematoria in Japan reported PCDDs/PCDFs
390 concentrations of 2.2-290 ng/N m³, giving toxicity equivalents (TEQ) of 0.0099-6.5 ng TEQ/N m³
391 (Takeda et al., 2000). In the same study, the estimated total PCDDs/PCDFs emission for all
392 crematoria in Japan was 8.9 gTEQ/year. PCBs have also been detected in concentrations of 16
393 ng/L (mean) and 290 ng/L (maximum) in embalming wastewater (Kleywegt et al., 2019).

394 PCDDs, PCDFs and PCBs have long residence times in the environment, hence they
395 undergo global transport over long distances and may have legacy health effects (Sonne et al.,
396 2010; AMAP, 2014). PCDDs, PCDFs and PCBs are lipophilic, hence bioaccumulate in the fatty
397 or lipid tissues of organisms, and undergo biomagnification along the trophic levels (Helgason et
398 al., 2013; Tartu et al., 2017; Kuchniski, 2019). Due to their carcinogenicity and high toxicity, they
399 pose significant health risks (Mari and Domingo, 2010; Kuckniski, 2019). A few reviews exist on
400 health risks of PCDDs/PCDFs in crematoria (Maria and Domingo, 2010; Cheruiyot et al., 2016).
401

402 **3.2.6 Illicit drugs**

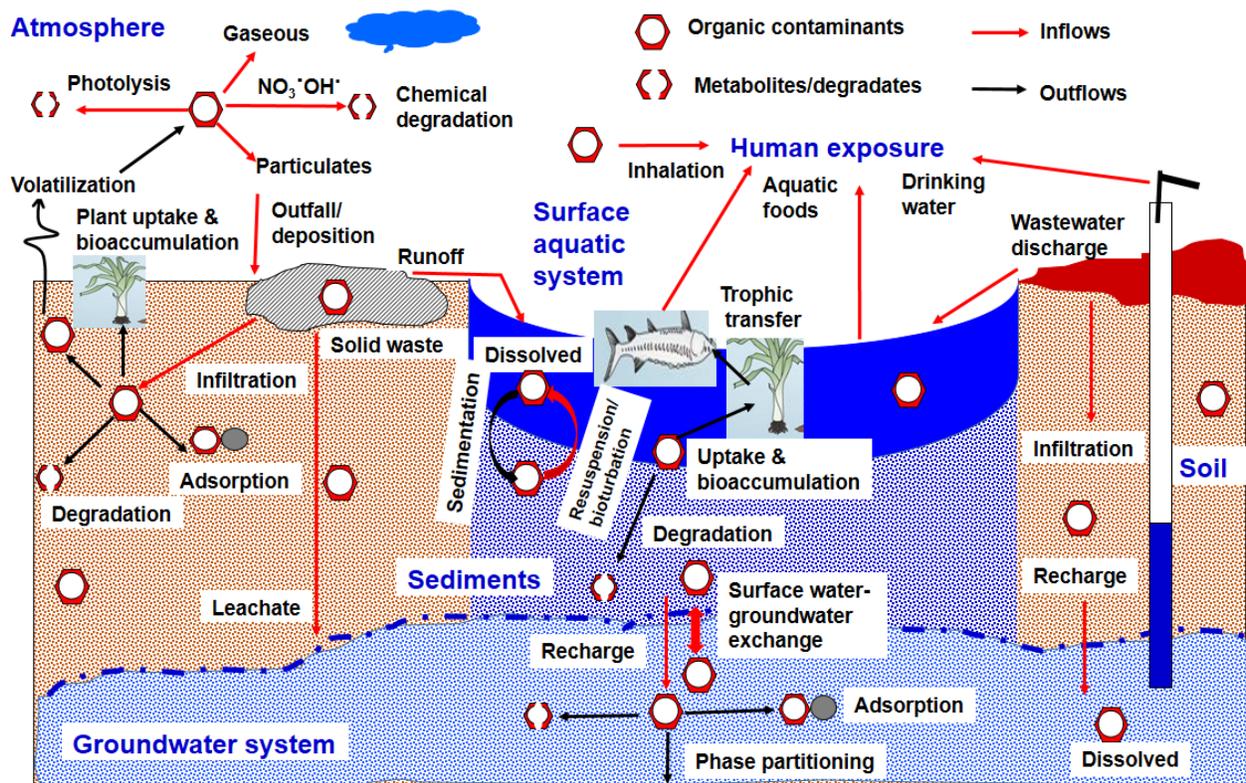
403
404 Data on illicit drugs are drawn mainly from autopsy and forensic studies (Jones and
405 Homgren, 2009; Pilgrim et al., 2011; Table 1). One exception is a study conducted in Ontario,
406 Canada, where high concentrations of Caffeine were detected in embalming wastewater, with
407 median and maximum values of 76650 ng/L and 4,200,000 ng/L, respectively (Kleywegt et al.,
408 2019). High concentrations of illicit drugs were also reported in post-mortem and forensic studies
409 (Pilgrim et al., 2011). A study based on the national database of the National Coroners Information
410 System in Australia showed that the concentrations of MDMA, also known as 'Ecstasy' in post-
411 mortems ranged from 0.02 to 3.5 mg/L. (Pilgrim et al., 2011). In the same study, other illicit drugs
412 detected in concentration ranges of mg/L were; Amphetamines, Codeine, Cocaine and Morphine.
413 A study of 49 autopsy cases covering a period of 11 years in Victoria (Australia) reported Cocaine
414 concentrations ranging from 0.01-3.0 mg/L with a median of 0.1 mg/L (Pilgrim et al., 2013). In
415 Sweden, analysis of 24, 876 forensic autopsies detected several illicit drugs and their metabolites
416 in femoral venous blood samples (Jones and Homgren 2009). Morphine and Codeine were
417 detected with mean concentrations of 0.3 mg/L and codeine 0.32 mg/L, respectively.
418

Notably, the concentrations of illicit drugs reported here for embalming wastewater and

419 autopsies are several orders of magnitude higher than the ng/L ranges reported in surface waters
 420 and wastewaters (Bones et al., 2007; Zuccato et al., 2005, 2008; Rosi-Marshall et al., 2015).
 421 However, the concentrations of several other common illicit drugs remain unknown. Besides
 422 autopsy and forensic studies, and embalming, data on the occurrence of illicit drugs in other
 423 environmental compartments such as surface and groundwater from cemeteries, solid wastes and
 424 air-borne particulates in the funeral industry are also scarce.

426 3.3 Dissemination, behaviour and fate

427
 428 Figure 2 presents a generalized summary of the dissemination, behaviour and fate of TOCs
 429 in the environment. A detailed discussion of the biogeochemical behaviour of organic
 430 contaminants, including emerging contaminants is presented in earlier reviews (Pal et al., 2010;
 431 Stuart and Lapworth, 2013; Gwenzi and Chaukura, 2019). Hydrological processes such as erosion,
 432 runoff, sub-surface flow, infiltration, recharge, and surface water-groundwater exchanges
 433 disseminate TOCs into the environment (Lunenberg et al., 2018; Gwenzi and Chaukura, 2018).
 434 TOCs, including volatiles such formaldehyde, disinfectants, perfumes and fragrances can also be
 435 disseminated via air-borne particulates, aerosols and gases (Asare-Donkor et al., 2020; Tratnyek
 436 et al., 2020). Hence, TOCs may ultimately occur in the atmosphere, terrestrial ecosystems (e.g.,
 437 soils), surface aquatic systems, groundwater systems, and biota, including humans (Figure 2).
 438
 439



440
 441
 442 **Figure 2.** A generalized conceptual depiction of the circulation, behaviour and fate of organic
 443 contaminants in the various environmental compartments
 444

445 TOCs undergo diverse biogeochemical behaviours, (bio)transformation and fate processes
446 (Pal et al., 2010; Lapworth et al., 2012; Stuart and Lapworth, 2013). The behaviour and fate depend
447 on biogeochemical conditions and the physico-chemical properties of the TOCs (Gwenzi and
448 Chaukura, 2018, Figure 2). Physico-chemical properties include; solubility, partitioning
449 coefficients, charge, lipophilicity and molecular diffusion coefficients (Stuart and Lapworth, 2012;
450 Shields et al., 2014). Biogeochemical conditions include redox potential, organic carbon, pH,
451 biochemical processes, and soil type (Shields et al., 2014). TOCs such as formaldehyde and POPs
452 may volatilize and accumulate into the atmosphere where they in turn, undergo partitioning among
453 gaseous, aerosols and adsorbed particulate phases (Figure 2). TOCs in the atmosphere may also
454 undergo photochemical degradation catalysed by OH and NO₃ radicals (Shields et al., 2014).

455 TOCs in soils undergo phase partitioning between dissolved and adsorbed phases, re-
456 volatilization, biochemical degradation, and uptake and bioaccumulation in biota (Fu et al., 2018;
457 Figure 2). In surface aquatic systems, TOCs may undergo the following processes (Gwenzi and
458 Chaukura, 2018): (1) sedimentation and re-suspension via bio-perturbation in bioactive zones, (2)
459 uptake and bioaccumulation by aquatic biota, (3) adsorption onto sediments, and (3) degradation
460 and re-volatilization in surface water layers (Figure 2). In groundwater systems, TOCs undergo
461 partitioning between bulk and pore water, dissolved and particulate phases, adsorption on the solid
462 matrix of the aquifer, and degradation (Drewes, 2003). However, further research involving the
463 application of partitioning, transport and speciation models is required to understand the long-term
464 behaviour and fate of TOCs especially in tropical environments.

465

466 **4 Exposure and health risks of TOCs**

467

468 **4.1 Human exposure pathways**

469

470 Human exposure to TOCs occurs via occupational and non-occupational exposure
471 (Nimmen et al., 2006). High-risk workers include those in autopsy, embalmers, funeral directors,
472 undertakers, grave diggers and cremators (Gwenzi, 2020). Exposure pathways include; (1) dermal
473 intake via bruises, cuts and wounds, and (2) oral route via faecal matter and body fluids during
474 dissection and manipulation of human cadavers (Davidson and Benjamin 2006). Air-borne
475 particulates are reservoirs of TOCs, including personal care products (Shin et al., 2020). Hence,
476 inhalation of air-borne particulates and aerosols may occur in autopsy, thanatopraxy, funeral
477 homes, cemeteries and crematoria. Inhalation is likely to occur for volatile TOCs such as
478 formaldehyde, some pesticides, aerosol antimicrobials and personal care products, and POPs
479 (Takeda et al., 2004; Aronson et al., 2006). Non-occupational exposure occurs via ingestion of
480 contaminated marine/aquatic foods (Llobet et al., 2008; Domingo and Bocio, 2017), and untreated
481 drinking water (Gwenzi, 2020), and inhalation of PCDDs, PCDFs and PCBs for communities
482 living close to crematoria (Takeda, 2000, 2004). Non-occupational exposure may also occur via
483 inhalation of volatile TOCs and dermal contact during body viewing, washing and handling of
484 human cadavers during home care, funerals and burials.

485

486 **4.2 Risk factors: an African perspective**

487

488 Earlier studies postulated that exposure and health risks of contaminants could be
489 particularly high in Africa and other developing regions (Gwenzi and Chaukura, 2018; Gwenzi,
490 2020). The funeral industry is no exception, hence a similar scenario is expected due to several

491 risk factors. Africa has weak and poorly enforced environmental and occupational health
492 regulations (Loewenson, 1995), and a thriving informal ‘black ‘market for toxic chemicals (e.g.,
493 DDT) and even pharmaceuticals (Schwarzenbach et al., 2006; Rother, 2010). A high vector and
494 disease burden necessitates frequent and widespread use and overuse of pesticides and
495 pharmaceuticals. The disposal of hazardous solid waste in non-engineered waste dumps, and
496 discharge of raw or partially treatment wastewaters into aquatic systems are common (Ali et al.,
497 2017; Angassa et al., 2020). Data on health risk assessments conducted prior to registration and
498 approval of toxic chemicals are scarce, partly because Africa largely relies on imports of chemicals
499 or formulations prepared in developed countries (Gwenzi and Chaukura, 2018). In this regard,
500 evidence on health risks from developed countries is assumed to be valid for the tropical
501 ecosystems in Africa. Rudimentary and unhygienic burial practices are common in some socio-
502 cultural and religious settings (Santarsiero et al., 2000; Zume, 2011; Ringane et al., 2017; Turajo
503 et al., 2019). This includes home burials, where graves are located close to shallow wells and
504 boreholes and even wetlands (Zume, 2011; Abia et al., 2019; Turajo et al., 2019). This poses risks
505 for groundwater contamination via cadaver slurries and leachates (Abia et al., 2019, Figure 2). Yet
506 a significant portion of the population in Africa relies on untreated drinking water from unsafe
507 surface water and shallow groundwater sources (Potgieter et al., 2020). In addition, low literacy
508 levels imply that communities often lack understanding of the health risks of TOCs in the funeral
509 industry. Therefore, understanding the health risks of TOCs in such high-risk regions is critical.

510

511 **4.3 Human and ecological health effects**

512

513 The health risks of TOCs depend on: (1) nature, concentration, and speciation of the
514 contaminant, and (2) the receptor organisms, including species, developmental stage and exposure
515 route, and (3) the co-occurrence of antecedent contaminants and health stressors. The health risks
516 of various groups of TOCs detected are presented in Table 2 and several articles (Kümmerer, 2010;
517 Boxall et al., 2012; Stuart et al., 2012; Vandenberg et al., 2012, William and Brooks, 2014; Liu
518 and Wong, 2013; Aronson et al., 2014; ATSDR, 2019; Kuchniski et al., 2019). In summary, the
519 health risks include: (1) endocrine disruption, (2) teratogenicity, (3) carcinogenicity, (4)
520 genotoxicity, (5) developmental toxicity, (6) cell and organ toxicity, (7) disruption of information
521 flow (info-disruption), and (8) behavioural changes (Table 2). However, the health risks are largely
522 based on inferential evidence because data directly relating the occurrence of specific TOCs in the
523 funeral industry to health outcomes are scarce. The only exception is formaldehyde, which has
524 been associated with health outcomes among embalmers (Aronson et al., 2004).

525

526

527

528

529

Table 2. Summary of human and ecological health risks of toxic organic contaminants reported in the funeral industry

Organic contaminants	Health risks and references	Remarks
A: Embalming products:		
(4) Formaldehyde/formalin	A highly volatile class 1 human carcinogen associated with various types of human cancers (IARC, 2006). Cancers reported in embalmers (Hauptmann et al., 2009; Aronson et al., 2014;	Human health risks high via occupational exposure (e.g., embalmers). Limited data on health effects on soil organisms
(5) Triclosan	Co-selects and promote antimicrobial resistance (Hartmann et al., 2016; Webber et al., 2017).	Data on risks suffers from methodological flaws (Goodman et al., 2018)
(6) Phenol	Acute and chronic toxicity in human and aquatic biota have been reported (Breton et al., 2003; Zhong et al., 2012).	Phenol has a short half-life e.g., 70 hr at pH 5-8.1 temperature 22-28°C
B: Persistent organic pollutants (dioxins, furans and PCBs):		
(3) Polychlorinated dibenzodioxins (PCDDs), polychlorinated dibenzofurans (PCDFs),	Carcinogenicity and defective neurodevelopment in infants in Europe and the USA (Arisawa et al., 2005). Disrupt endocrine and reproductive systems in arctic species (Sonne et al., 2017). Reduced testosterone and sperm viability in Arctic foxes (<i>Vulpes lagopus</i>) (Sonne et al., 2017). Decreased levels of steroids in polar bears (<i>Ursus maritimus</i>) (Gustavson et al., 2015).	Data on other human health effects inconsistent or has lack sufficient statistical power (Arisawa et al., 2005).
C: Synthetic organic pesticides:		
(1) Methyl parathion	Neurotoxic via inhibition of acetylcholine esterase in nervous systems, genotoxic and carcinogenic (Mulla et al., 2020)	Behaviour and health risks discussed in recent reviews (Mulla et al., 2020).
(2) Imidacloprid [1-(6-chloro-3pyridylmethyl)-N-nitroimidazolidin-2-ylideneamine]	Acute toxicities in humans, and oxidative stress causing reproductive disorders including oestrous cycle defects and infertility in terrestrial and aquatic organisms (Arya et al., 2019).	Human fatal toxicities reported in cases of suicide (Proença et al., 2005).
(3) Methomyl (S-methyl-N-[(methylcarbamoyl)oxy]thioacetimidate)	Oxidative stress and acute toxicity reported in humans and biota. Sub-lethal concentrations alter feeding habits, and disrupt endocrine and reproductive system in avians (Arya et al., 2019).	Human health risks such as fatal poisoning associated with suicide (Tsatsakis and Tsakalof, 1996).
(4) Lindane	Highly persistent and toxic, and listed in the Stockholm Convention (http://chm.pops.int/TheConvention/ThePOPs/ListingofPOPs/tabid/2509/Default.aspx)	Banned in most countries, but informal markets exist in developing countries.
(5) pp-DDE (1,1-dichloro-2,2-bis(p-chlorophenyl) ethylene)	Obesity and metabolic disruption via inhibition of electron transport and oxidative phosphorylation (Elmore and La Merrill, 2019). Adverse effects on human sperm and alteration of the sperm Y: X chromosome ratio reported in three European populations and Inuit population In Greenland (Tiido et al., 2006)	ATSDR (2019) presents a recent review on health risks of DDT and its metabolites

532 **Table 2** (contd.)

533

Organic contaminants	Health risks and references	Remarks
<i>E: Pharmaceuticals and illicit drugs:</i>		
(4) Pharmaceuticals	Promote the development of antimicrobial resistance (Gullberg, 2014). Wide range of health effects reported for various pharmaceuticals drugs (Li et al., 2020; Sathishkumar et al., 2020).	Reviews exist on health risk of pharmaceuticals (Li et al., 2020; Sathishkumar et al., 2020)
(5) Illicit drugs	'Pseudo-persistent due to frequent use. Concentration-dependent acute and chronic toxicities may occur in human and biota (Rosi-Marshall et al., 2015). Aquatic organisms such as bacteria, fish, invertebrates have receptors hence could be sensitive (Rosi-Marshall et al., 2015).	Human health effects including fatalities limited to deliberate intake via drug abuse (Pilgrim et al., 2011). Further research on ecological effects required in high use areas.
<i>F: Personal care products:</i>		
(1) Oxybenzone	Concentration-dependent toxicity and coral bleaching in laboratory studies (Adler and DeLeo, 2020). Malformation, teratogenicity and neurodegenerative effects in tadpoles (<i>Bufo arabicus</i>) exposed to 10 ppm (Seleem et al., 2018)	Epidemiological evidence linked to human health outcomes remain weak.
(2) DEET (N,N-Diethyl-m-toluamide)	Ecotoxicological data from China showed that DEET had higher toxicity in aquatic algae than animals (Gao et al., 2020).	Endocrine disruption and neurotoxicity reported in other personal care products (Shin et al., 2019).
(3) Nonyl phenol and its metabolites	Endocrine disruption via sex steroid hormone receptors and conversion of testosterone to oestrogen. They also act as info-disruptors (Kuzikova et al., 2019).	Several personal care products cause info-disruption (Parrish et al., 2019)

534

535

536 4.3.1 Pharmaceuticals

537
538 Pharmaceuticals cause a wide range of ecological health effects (Li et al., 2020;
539 Sathishkumar et al., 2020, Table 2). The development of antimicrobial resistance is the most cited
540 health risk of pharmaceuticals and their metabolites (Hughes and Andersson, 2012).
541 Pharmaceuticals have low ecotoxicological thresholds, and can induce adverse health effects even
542 at low concentrations (e.g., ng/L) (Paredes et al., 2014). This is because pharmaceuticals are
543 designed to have therapeutic effects even at low doses (Paredes et al., 2014). Hence, even sub-
544 lethal concentrations induce antimicrobial resistance in susceptible organisms (Hughes and
545 Andersson, 2012; Jutkina et al., 2016). Antimicrobial resistance increases pathogenicity and
546 promote outbreaks of antimicrobial resistant infections. For example, Aspirin, Ibuprofen and
547 Diclofenac induce bacterial resistance to antibiotics, and sterility and feminization in aquatic
548 animals (Agunbiade and Moodley, 2014). Some pharmaceuticals are endocrine disruptors that alter
549 the hormonal system via blocking and biomimicry (Bolong et al., 2009). Disruption of trophic
550 interactions and chemical information flows (info-disruption) have been reported in aquatic
551 systems (Van Donk et al., 2016).

552 Studies have investigated the health risks of human exposure to pharmaceuticals via
553 aquatic and marine foods, drinking water, and air-borne particulates. Results for the period up to
554 2010 concluded that human exposure to pharmaceuticals via these exposure pathways posed low
555 to no appreciable human health risks (Schulman et al., 2002; Webb et al., 2003; Fent et al., 2006;
556 Nimmen et al., 2006; Kummerer, 2008; Bottoni et al., 2010; Kumar et al., 2010). Subsequent
557 studies covering the period from 2011 to 2019 confirmed the previous findings, and concluded that
558 there is no or low risk to human health for all age groups (Houtman et al., 2014; Prosser and
559 Sibley, 2015; Paltiel et al., 2016; Li et al., 2017; Tomasi et al., 2017; Brown et al., 2018; Fantuzzi
560 et al., 2018; Semerjian et al., 2018; Kibuye et al., 2019; Letsinger and Kay, 2019; Praveena et al.,
561 2019; Sharma et al., 2019). The only exceptions are a few studies suggesting health risks in
562 children and infants consuming crops or tap water contaminated with pharmaceuticals (Leung et
563 al., 2013; Malchi et al., 2014). These conclusions are consistent with that of the WHO (2012),
564 who concluded that no adverse human health effects exist due to chronic exposure to
565 pharmaceuticals in drinking water. Other authors are more cautious: Kümmerer (2010) concluded
566 that the short-term effects of pharmaceuticals in humans are not known, while Tourand et al. (2011)
567 concluded that no consensus exists on the topic.

568 However, current evidence on human health risks has several inherent limitations and
569 assumptions (*Section 4.4*). For example, only single exposure routes to parent pharmaceuticals
570 were considered, while multiple co-exposures and equally toxic metabolites were excluded.
571 Interactions among pharmaceuticals, and other health stressors were not considered. Literature is
572 also silent on the health risks associated with long-term human exposure to sub-lethal
573 concentrations, yet pharmaceuticals are pseudo-persistent. The bulk of studies on human health
574 risks were limited to developed countries (Tomasi et al., 2016; Brown et al., 2019), while no
575 evidence exists in developing regions including Africa. Yet the occurrence and health risks of
576 pharmaceuticals in the environment is global concern (Kurster and Adler, 2014).

577 578 4.3.2 Embalming products

579
580 Formaldehyde and phenols are well-known human carcinogens (IARC, 2006; Table 2).
581 Significant associations have been established between formaldehyde exposure and high

582 incidences of certain cancers among embalmers (Hauptmann et al., 2009; Aronson et al., 2014). A
583 review of retrospective studies investigating the association between formaldehyde and
584 teratogenicity or adverse developmental and reproductive effects showed that women exposure to
585 formaldehyde increased the risk of incidences of spontaneous abortion (Duong et al., 2011). Acute
586 and chronic human toxicity due to phenol exposure also occur (Breton et al., 2003). A case-control
587 study of a glutaraldehyde-based organic glue suggest that glutaraldehyde may induce intense
588 inflammation, causing wounds to breach, thereby promoting the proliferation and spreading of
589 bacteria (Gaberel et al., 2011).

590 The ecological effects of formaldehyde and phenol include acute and chronic toxicity
591 (Breton et al., 2003; Zhong et al., 2012). Phenol and its metabolites cause endocrine disruption in
592 terrestrial and aquatic animals (Breton et al., 2003; Zhong et al., 2012). Phenol toxicity depends
593 on the developmental stage of the receptor organisms. The embryo-larval stage is the most
594 sensitive in fish such as rainbow trout (*Oncorhynchus mykiss*), and amphibians such as leopard
595 frog (*Rana pipiens*) (Breton et al., 2003). However, the health risks of other TOCs in embalming
596 products such as penetrating agents, moisturizers and anti-coagulants remain unknown.

597

598 **4.3.3 Personal care products**

599

600 Compared to other personal care products in the funeral industry, Triclosan is one of the
601 most studied with respect to human health risks, but the results are mixed (Clayton et al., 2011;
602 Savage et al., 2011; Cullinan et al., 2012; Aronson et al., 2014). On one hand, oral exposure to
603 Triclosan reduced the concentration of serum thyroid hormone in test animals (Aronson et al.,
604 2014). This is because the structure of Triclosan shares some similarities with thyroid hormones.
605 On the other hand, a controlled randomized study of 132 subjects including a placebo showed no
606 significant differences in thyroid function and antithyroid antibodies between the control groups
607 and those exposed to a toothpaste containing 0.3% triclosan for a year (Cullinan et al., 2012).
608 Savage et al. (2011) investigated a case study of a 50-year old kitchen worker, who used a
609 Triclosan-containing aerosol disinfectant. The subject had developed adverse health conditions
610 including persistent facial eruptions, gross periorbital edema, facial erythema and pruritus, which
611 showed irregular appearance and clearing, and often involved the chest, upper shoulders and neck
612 (Savage et al., 2011). Results of patch tests using a baseline (control), facial and hair-dressing
613 formulations showed a positive reaction to 2% Triclosan in the petrolatum (Savage et al., 2011).
614 In USA, a study analysed national data obtained from the National Health and Nutrition
615 Examination Survey for the period 2003 to 2006 (Clayton et al., 2011). The results showed that
616 higher urinary concentrations of triclosan were significantly associated with increased incidences
617 of allergies such as hay fever (Clayton et al., 2011). Other studies show that Triclosan co-selects
618 and promotes antimicrobial resistance (Hartmann et al., 2016; Webber et al., 2017), but studies
619 investigating this aspect in humans are still lacking.

620 Table 2 presents the ecological health risks of the detected personal care products.
621 Oxybenzone caused concentration-dependent toxicity and coral bleaching in laboratory studies
622 (Adler and DeLeo, 2020). Malformation, teratogenicity and neurodegenerative effects were
623 reported in tadpoles (*Bufo arabicus*) exposed to 10 ppm (Seleem et al., 2018). Ecotoxicological
624 data from China showed that DEET had adverse toxicity on aquatic organisms, and the toxicity
625 effects were higher in aquatic algae than animals (Gao et al., 2020). Endocrine disruption via sex
626 steroid hormone receptors and conversion of testosterone to oestrogen have been observed for
627 nonyl phenol and its metabolites. Several personal care products are info-disruptors via

628 biomimicry of natural info-chemicals such as pheromones (Kuzikova et al., 2019). Info-disruption
629 may alter reproductive behaviour and trophic interactions. Further research is required to
630 understand the health risks of other personal care products commonly used in the funeral industry.
631

632 **4.3.4 Synthetic organic pesticides**

633

634 Post-mortem and forensic studies report the acute and chronic toxicity, and human fatalities
635 caused by synthetic organic pesticides including organophosphates and carbamates (Tsatsakis and
636 Tsakalof 1996; Proenca et al., 2005; Table 2). Organophosphates such as Methyl Parathion are
637 genotoxic, carcinogenic and neurotoxic due to their ability to inhibit acetylcholine esterase in
638 nervous systems (Mulla et al., 2020). Genotoxicity via sperm chromatin alteration and DNA
639 damage caused by oxidative stress have been reported (Salazar-Arredondo et al., 2008).
640 Imidacloprid and Methomyl cause acute toxicity, oxidative stress and reproductive disorders
641 including oestrous cycle defects (Proenca et al., 2005; Arya et al., 2019). DDE causes fertility
642 disorders, including the alteration of the sperm Y: X chromosome ratio (Tiido et al., 2006; Taylor
643 et al., 2007; Dallaire et al., 2008; Elmore and Merrill, 2019).

644 The ecological health risks of pesticides is one of the most studied among TOCs (Table 2).
645 Concentration-dependent chronic and acute toxicity of DDE, including tremors with death,
646 reproductive disorders (e.g. reduced fertility), neurodevelopmental changes, and endocrine
647 disruption (ATSDR, 2019) have been reported in animals (ATSDR, 2019). Imidacloprid, Methyl
648 Parathion and Methomyl cause reproductive disorders, endocrine disruption and oxidative stress
649 in a number of avian and animal species (Arya et al., 2019; Mulla et al., 2020, Table 2). For
650 example, sub-lethal concentrations of Methomyl alter feeding habits, and disrupt endocrine and
651 reproductive system in avian and aquatic species (Arya et al., 2019).

652 Besides post-mortem and forensic studies, the health risk of pesticides in the funeral
653 industry have received limited research attention. The notion that homicide and suicide cases are
654 relatively rare compared to natural deaths due to infectious diseases may explain this trend.
655 However, data show that homicide and suicide cases are prevalent in developing countries with a
656 thriving agricultural sector (Gunnell and Eddleston, 2003). Further research is needed to
657 understand the occupational and non-occupational health risks of pesticides in the funeral industry.
658

659 **4.3.5 Persistent organic pollutants**

660

661 PCDDs, PCDFs and PCBs have been detected in human foods including meat and edible
662 aquatic and marine foods including fish, ambient air and air-borne particulates (Libet et al., 2008;
663 Domingo and Bocio, 2017). Hence, human exposure occurs via: (1) ingestion or dietary intake of
664 contaminated marine and aquatic foods, and (2) inhalation of air-borne particle such as soot and
665 dust (Libet et al., 2008; Domingo and Bocio, 2017). PCDDs, PCDFs and PCBs pose carcinogenic
666 or cancer risks in humans (Xu et al., 2020), and adverse human health risks have been reported in
667 terms of pregnancy outcomes (Dummer et al., 2003). Human exposure to PCDDs and PCDFs has
668 been linked to defective neurodevelopment in infants in Europe and the USA (Arisawa et al.,
669 2005). One study showed that PCDD/PCDF concentrations of 7.00–215 pg/g in soils from point
670 sources such as smelters may result in cancer risks of 0.487×10^{-6} (Xu et al., 2020).

671 The occurrence and health risks of PCDDs, PCDFs and PCBs have been detected in
672 terrestrial and aquatic biota including avian, aquatic and marine species (Dallaire et al., 2006,
673 Kwon et al., 2019; Lehikoinene et al., 2019, Table 2). Studies have reported the adverse health

674 effects of dioxins and furans on reproductive and hormonal systems in Arctic polar bears (*Ursus*
675 *maritimus*), Arctic foxes (*Alopex lagopus*, *Vulpes lagopus*) and sled dogs (*Canis lupus familiaris*)
676 (Gustavson et al., 2015; Muijten et al., 2016; Riget et al., 2016, 2019; Sone et al., 2017). In these
677 species, POPs reduced testosterone and sperm viability in Arctic foxes (*Vulpes lagopus*) (Sonne et
678 al., 2017), and concentration of circulating steroids in female polar bears (*Ursus maritimus*)
679 (Gustavson et al., 2015).

680 Data on the health effects on species occurring tropical ecosystems in tropical
681 environments particularly in Africa remain limited. Further research is also required to determine
682 the occurrence of PCDDS, PCDFs and PDBS in human media such as hair, nails and body fluids
683 among crematorium workers and surrounding communities. Particular attention should be paid to
684 female workers in the cremation industry and their children due to the potential risk of transfer of
685 POPs from mothers to their unborn fetuses and infants.

686

687 **4.3.6 Illicit drugs**

688

689 Evidence on human health risks of illicit drugs is limited to acute and chronic toxicity, and
690 fatalities due to drug abuse (Pilgrim et al., 2011). Illicit drugs induce behavioural and physiological
691 changes in humans (Rosi-Marshall et al., 2015). Illicit drugs containing plant-derived alkaloids
692 may have antimicrobial properties (Radulovic et al., 2013), hence oral intake of such drugs may
693 alter the microbial diversity and function in the human mouth and gut. For example, significant
694 differences in microbial diversity were observed between humans chewing *khat*, a plant containing
695 stimulants resembling monamine amphetamine compared to non-khat chewers (control) (Al-
696 Hebshi et al., 2010). However, it is currently unclear whether human exposure to environmentally
697 relevant concentrations of illicit drugs via inhalation and ingestion cause adverse health effects.

698 The ecological health risks of illicit drugs have been reported in a number of studies,
699 including reviews (Binelli et al., 2012; Parolini and Binelli, 2013, 2014; Rosi-Marshall et al.,
700 2015). Aquatic organisms such as bacteria, fish and invertebrates have receptors that make them
701 sensitive to illicit drugs (Rosi-Marshall et al., 2015). The antimicrobial properties of some illicit
702 drugs have been reported to disrupt microbial community composition and function (Radulovic et
703 al., 2013). Amphetamines disrupt catecholamine production and reception in mammals and algae,
704 thereby disrupting ecological function and interactions (Rosi-Marshall et al., 2015). Exposing
705 zebra mussel, an aquatic organism to Cocaine, and its metabolites at environmentally relevant
706 concentrations caused adverse physiological and genotoxicity effects including DNA damage via
707 oxidative stress (Binelli et al., 2012; Parolini and Binelli, 2013, 2014; Parolini et al., 2013).
708 However, current evidence on ecological health risks only cover a few illicit drugs and species,
709 while a large number of other illicit drugs and species remain under-studied. Thus, the health
710 effects of illicit drugs require further investigation.

711

712 **4.4 Health risks: A critique of the evidence and approaches**

713

714 The current review posed the question whether TOCs in the funeral industry constitute a
715 health risk or is a myth. The preceding discussion (**Section 4.3**), in conjunction with the occurrence
716 of TOCs (Table 1) and their health risks (Table 2) partly addressed this question. However, barring
717 formaldehyde exposure linked to health outcomes among embalmers (Aronson et al., 2006;
718 Hauptmann et al., 2009, limited evidence exists linking TOCs in the funeral industry to health
719 outcomes. This reflects the fact that, linking the occurrence of TOCs in environmental media to

720 health risks is not a trivial task. Yet the lack of evidence may be misinterpreted as lack of health
721 risks, thereby creating a technical loophole that can be exploited by unscrupulous players in the
722 chemical industry. The lack of direct evidence is not unique to TOCs in the funeral industry, but
723 is generic to other contaminants and industries, including antimicrobial resistance in drinking
724 water systems (Sanyangando and Gwenzi, 2019). Here, a few generic limitations and challenges
725 are highlighted to illustrate the difficulties and complexity of assessing health risks of TOCs.

726

727 **4.4.1 Methodological limitations**

728

729 Methodological limitations, inconsistencies among results and lack of sufficient statistical
730 power partly account for the lack of evidence. One study is illustrative in this regard: Arisawa et
731 al. (2005) reviewed global epidemiological studies linking human intake of dioxins, furans and
732 PCBs to diabetes mellitus, endometriosis, thyroid function and neurodevelopmental of infants.
733 Consistent results were only observed for neurodevelopmental disorders in infants which were
734 linked to human exposure to background concentrations of dioxins/PCBs. However, the
735 association between human exposure to dioxins, furans and PCBs, and diabetes, endometriosis and
736 thyroid function could not be established. This was due to: (1) lack longitudinal studies in the case
737 of diabetes, (2) inconsistent results for thyroid function, and (3) insufficient statistical power for
738 endometriosis. Other limitations include the use of methods that do not allow direct or quantitative
739 estimation of health risks. The bulk of studies merely reporting occurrence of TOCs (e.g., Fielder
740 et al., 2017; Kleywegt et al., 2019) without the corresponding data on bioaccessibility,
741 bioavailability and intake by target organisms fall under this category. In such cases, occurrence is
742 often misinterpreted as synonymous with adverse health effects. Some studies (e.g., case reports)
743 often lack proper statistical experimental design and data analysis, while others may lack sufficient
744 statistical power because of low sample size or limited replication. Hence, drawing definitive
745 conclusions from such studies could be problematic, and may even lead to misleading
746 interpretations. Therefore, future research should address these methodological flaws in existing
747 literature, and quantitatively determine the health risks of TOCs.

748

749 **4.4.2 The paradox of contaminant mixtures**

750

751 Health risk assessment protocols largely rely on *in vitro* and *in vivo* bioassay experiments
752 (e.g., EC, 2006; OECD, 2015; US EPA, 2020a, b). This often involves exposing target bioassay
753 organisms to a single chemical at a given concentration via a specific exposure, followed by
754 qualitative or quantitative estimation of risk (Linov et al., 2001). Such bioassay tests often use
755 ‘benign’ media spiked with target contaminants (OECD, 2015; EC, 2006), while excluding
756 interactions and health effects of antecedent contaminants and health stressors. In real systems,
757 receptor organisms including humans are simultaneously exposed to a plethora of TOCs and co-
758 occurring health stressors at various levels via multiple exposure routes. Co-exposure to mixtures
759 of TOCs, and health stressors change the exposure scenario, resulting in complex dose-response
760 behaviours. Such complex interactions include additive, antagonistic, neutral and synergistic
761 health effects (Carvalho et al., 2014; Oskarsson et al., 2014; Chen et al., 2017; Parrish et al., 2019).

762 Chen et al. (2017) showed that, depending on specific toxicological endpoints, synergistic
763 and antagonistic interactions between Pb and decabromodiphenyl ether occurred concurrently. A
764 study involving 16 laboratories in Europe investigated the effects of mixtures of 14 or 19 chemicals
765 drawn from various groups using 35 bioassay tests involving 11 organisms from various trophic

766 levels (Carvalho et al., 2014). The concentrations of all chemicals tested were set at the safety
767 limits prescribed by the European Community regulations. Adverse health effects of mixtures were
768 observed, including microalgae toxicity, fish embryo toxicity, embryo developmental toxicity and
769 oxidative stress (Carvalho et al., 2014). In a binary mixture of two fungicides (Prochloraz,
770 Azoxystrobin), the no-observed-effect concentration of Prochloraz reduced the median effect
771 concentration of azoxystrobin in *Hyaella Azteca* and *Gammarus pulex*) (Fu et al., 2018).

772 The health risks of mixtures pose potential regulatory and research challenges. First, each
773 single synthetic chemical is designed to have a particular effect/s at a particular dose. Therefore,
774 although health risks of mixtures are closer to reality than single contaminants, it will be
775 problematic to base regulatory approval decisions on data considering the health effects of a suite
776 of contaminants. The mechanisms of how the interactions among TOCs, and with antecedent
777 contaminants and health stressors increase toxicity remain unclear. In addition, whether or not
778 individuals, populations, communities and ecosystems co-exposed to mixtures of TOCs and other
779 health stressors exhibit physiological and phenotypic plasticity is poorly understood. A few studies
780 suggest that, some organisms may exhibit phenotypic plasticity in response to health stressors such
781 as those induced by climate change (Whiteman et al., 2018). Assuming that such ecological
782 plasticity occurs, there is need to understand the concentration ranges over which plasticity occurs,
783 beyond which the onset of adverse effects are expected. Further, Hayes et al. (2006) conjectured
784 that, contaminants in mixtures may act as effectors, enhancers or even as neutrals. Therefore, it is
785 equally important to determine the role of individual contaminants in mixtures of TOCs.

786

787 **4.4.3 Trans-generational, delayed and legacy effects**

788

789 Persistent and pseudo-persistent contaminants may also have trans-generational, delayed
790 and legacy health effects, which may not be detected in typical short-term ecotoxicological studies
791 (Tiido et al., 2006; Chen et al., 2017; Letcher et al., 2018). The developmental origins of health
792 and disease (DOHaD) concept suggests that human exposure to contaminants may lead to inter-
793 generational health effects (Lønnebotn et al., 2018; Miller and Lawrence, 2018). Thus, pre-
794 conception and early childhood exposure to TOCs may later determine human health outcomes as
795 reported for obesity, asthma and neurodevelopmental disorders (Lønnebotn et al., 2018; Miller and
796 Lawrence, 2018). The DOHaD concept could be relevant to persistent organic pollutants (dioxins,
797 furans, PCBs) and pesticides (e.g., DDT, DDE) (Dallaire et al., 2008; Tiido et al., 2006). Tiido et
798 al. (2006) report trans-generational effects of DDE on fertility, including the alteration of the sperm
799 Y: X chromosome ratio in three European populations and Inuit population (Table 2). Therefore,
800 inter-generational, delayed and legacy health effects of TOCs warrant further research.

801

802 **4.4.4 ‘Priority contaminants and the ‘equifinality’ concept**

803

804 There is an increasing interest to develop and use ‘lists of priority contaminants’ for
805 environmental surveillance and control systems (Bottoni et al., 2010; Li et al., 2020). Priority lists
806 may create an unjustified reification, a phenomenon whereby concepts, in this case the use of ‘lists
807 of priority contaminants’ become increasingly interpreted as scientific facts (Posthuma et al.,
808 2018). Besides influencing research funding decisions and choice of contaminants to monitor,
809 reification may lead to type I and II errors (Hyman, 2010; Posthuma et al., 2018). Type I errors
810 occur when misinterpretation of risk triggers implementation of mitigation measures, when in fact,
811 the risk is not related to true ecological or human health impacts (Prato et al. 2014). Type II errors,

812 occur when the health risks of several TOCs and their interactions in mixtures, and other health
813 stressors are unknown or neglected due imperfect knowledge caused by limitations of current
814 scientific methods. This is particularly the case for most emerging contaminants and their mixtures,
815 because their modes of action are still poorly understood.

816 Table 2 shows that several POPs, pharmaceuticals, and personal care products are
817 carcinogens, endocrine disruptors and info-disruptors. Thus, several TOCs may result in similar
818 health outcomes (Aronson et al., 2006; Rosi-Marshall et al., 2014). This phenomenon can be
819 summed up as the ‘equifinality’ concept, a term borrowed from the science of complex systems
820 (Aron, 2020). In the current context, the ‘equifinality’ concept implies that co-occurring TOCs
821 may not necessarily result in unique health outcomes. Instead, a single health outcome may arise
822 from several plausible TOCs. The ‘equifinality’ concept implies that, using conventional tools in
823 (eco)toxicology to partition the health effects among several co-existing TOCs presents
824 methodological challenges. Taken together, the concepts of ‘equifinality’ and the ‘priority
825 contaminants’ imply that, for contaminants with the same health effects, some may appear in the
826 ‘priority lists’ while others are excluded, hence they will be regulated differently. It is currently
827 unclear how type I and II errors, and the ‘equifinality’ concept are addressed in current health risk
828 assessment protocols.

829

830 **5 Assessment and mitigation of health risks**

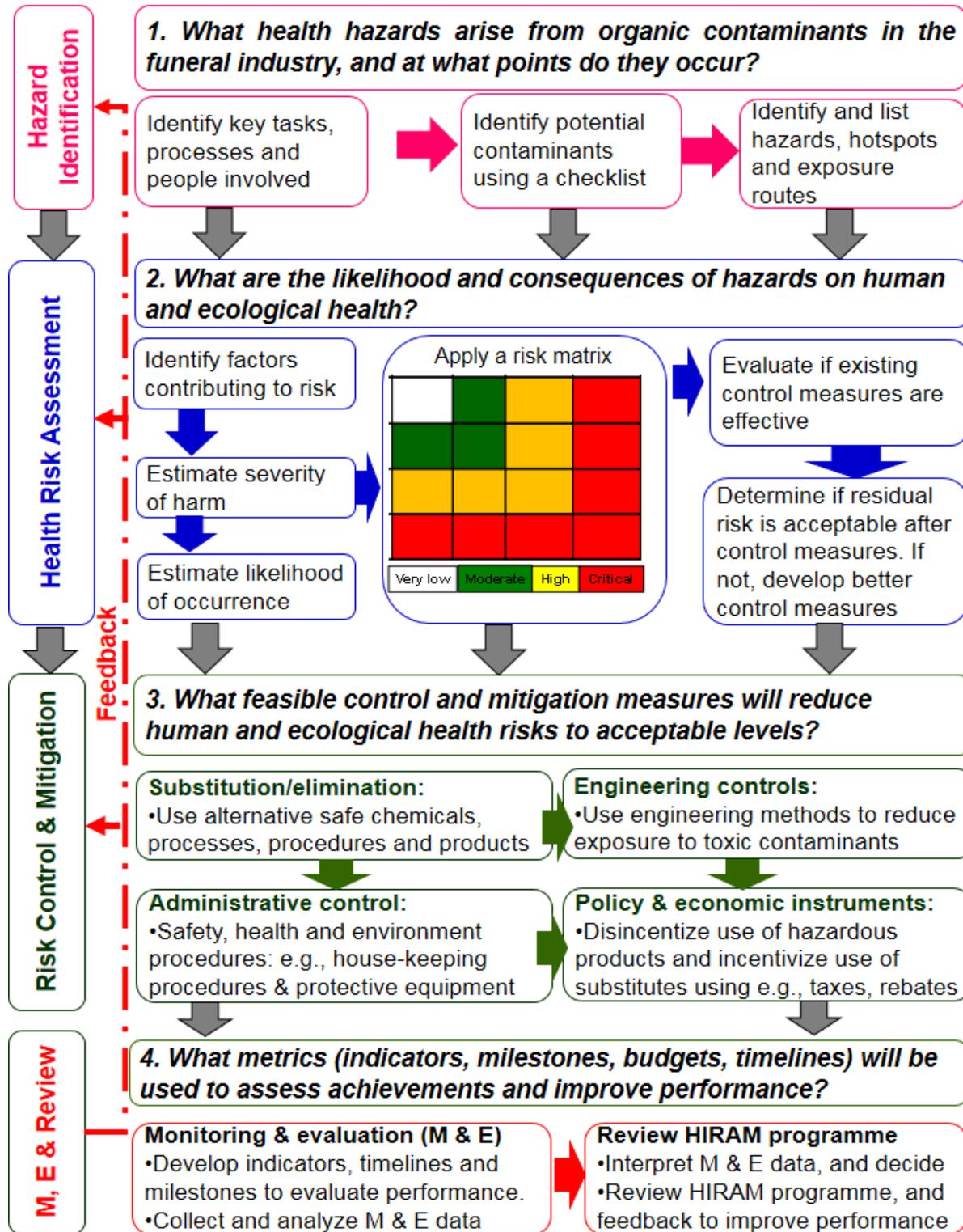
831

832 **5.1 A framework for health risk assessment and mitigation**

833

834 The overall goal of risk assessment and mitigation is to ensure that the health risks are
835 understood, communicated and mitigated to unacceptable levels. Figure 3 depicts a framework for
836 hazard identification, risk assessment, and mitigation (HIRAM). The HIRAM comprises of four
837 key steps: (1) hazard identification, (2) risk assessment, (3) risk mitigation and control, and (4)
838 monitoring and evaluation (M &E), and subsequent feedback. The key questions to guide each
839 step, and typical activities to be undertaken are also highlighted.

840



841
842
843

Figure 3. A conceptual framework for hazard identification, risk assessment and mitigation (HIRAM) in the funeral industry.

844 **5.1.1 Hazard identification**

845 This step characterizes the hazards posed by TOCs in key tasks and processes in the funeral
846 industry. This include; (1) determining the nature, sources and concentrations of TOCs, (2)
847 identifying people at risk, including embalmers, funeral directors, cleaners, assistants and service
848 providers, and (3) identifying key exposure pathways, and intakes per unit time (e.g. day).

849
850 **5.1.2 Risk assessment**

851 This entails qualitative and quantitative estimation of the likelihood or probability of
852 occurrence of a hazard, and the corresponding magnitude of consequences on health. Qualitative
853 risk analysis may use a matrix with qualitative terms such as ‘critical’, ‘high’, ‘moderate’,
854 ‘low/negligible’ to rate the likelihood of occurrence and magnitude of consequences (Figure 3).
855 Quantitative risk assessment calculates a risk metric or quotient (US EPA, 2020a, b) using tools
856 such as probabilistic techniques and quantitative structure-activity relationships (QSARs) (Walker
857 et al., 2002; Cronin et al., 2003). The estimated risk is then compared to set threshold values which
858 are considered to be acceptable. In the case of emerging contaminants, a key challenge is the lack
859 of established ecotoxicological threshold values required to induce adverse health effects. To
860 overcome this challenge, Gwenzi (2020) proposed a heuristic approach, whereby data from
861 pristine or non-impacted environments are used to provide the baseline or acceptable value.

862
863 **5.1.3 Risk mitigation and control**

864 This entails the identification and evaluation of the risk mitigation and control measures
865 based on results of risk assessment. The ultimate mitigation strategy may entail a combination of
866 the following: (1) administrative controls or ‘soft engineering’, (2) substitution and/or elimination
867 of hazardous chemicals, processes and procedures, (3) engineering controls to reduce exposure to
868 hazards, and (4) the use of regulatory and policy instruments (**Section 5.2**).

869 **5.1.4 Monitoring, evaluation and feedback**

870 This involves the development of appropriate indicators, timelines and milestones for
871 performance evaluation (Figure 3). Indicators may include fatalities and near-misses traced to
872 occupational exposure to hazards, and even frequency of occurrence of unsafe practices.
873 Milestones may include setting targets to reduce certain undesirable events by a specific period.
874 Subsequent analysis and interpretation of the M & E data will be used for decision-making,
875 including review and improvement of the HIRAM via a feedback process (Figure 3).

876
877 **5.2 Mitigation and control measures**

878
879 **5.2.1 Administrative measures**

880 These include measures often included in occupational safety, health and environment
881 procedures of organizations (Figure 3). Examples include; (1) practising good house-keeping
882 practices such as proper handling and disposal of hazardous chemicals and wastes, (2) use of
883 personal protective equipment, and (3) routine medical examinations and use of prophylactics. The
884 Globally Harmonized System (GHS) of Classification and Labelling of hazardous materials and
885 material safety datasheets (MSDS) are often used for hazard/risk communication, and proper
886 labelling and signage to indicate the existence and nature of hazards (Winder et al., 2005; US
887 OSHA, 2013). Other options include training workshops, and awareness and educational
888 campaigns targeting key stakeholders such as occupational workers, environmental regulators, and

889 environmental and public health officers. Routine occupational and environmental surveillance
890 systems are critical for the early detection and remediation of health risks (Loewenson, 1995).

891

892 **5.2.2 Substitution and elimination**

893 This entails the substitution, replacement and even elimination of hazardous materials,
894 processes and procedures. For example, hazardous embalming chemicals based on formaldehyde
895 can be substituted or replaced by several newly developed and less hazardous alternatives (Varlet
896 et al., 2019). Even the embalming process can be eliminated in cases where it is not warranted. In
897 some societies, this may entail a shift from current unsafe funeral practices such as home burials
898 (Zume, 2011; Turajo et al., 2019) to safer alternative practices that pose limited health risks. In
899 such societies, this may require long-term changes in certain socio-cultural and religious beliefs.

900 Emerging technologies for the disposal of human cadavers such as freeze-drying and
901 alkaline hydrolysis can also be used as alternatives for embalming and subsequent burial in
902 cemeteries, and cremation. Da Cruz et al. (2017) presents a detailed discussion of the various
903 funeral technologies including the relative advantages of the emerging technologies. Freeze-drying
904 or promession entails subjecting whole human cadavers to rapid freezing using liquid nitrogen,
905 followed by burial without embalming (Da Cruz et al., 2017). Alkaline hydrolysis involves subject
906 a cadaver to concentrated potassium hydroxide (KOH) and steam at a temperature of 180°C for a
907 few hours, and creating a whirlpool effect to accelerate dissolution (Da Cruz et al., 2017). The end-
908 products of alkaline hydrolysis are liquid containing dissolved tissues, and bones. The bones are
909 turned into a powder using a cremulator or crusher and given to the bereaved family for storage,
910 while the liquid can be used as a fertilizer (Everts et al., 2010). The potential benefits of these
911 emerging technologies relative to conventional ones include: (1) low land requirement and costs,
912 estimated to be about US\$389 for freeze-drying compared to US\$423 for cremation and US\$467-
913 622 for burial, and (2) low environmental footprints including reduced contamination by organic,
914 inorganic and microbiological contaminants, and low energy requirement compared to cremation
915 (Da Cruz et al., 2017). However, the costs of initial installation and operation of emerging
916 technologies could be prohibitive especially in developing countries. Moreover, quantitative data
917 on the environmental benefits and the long-term behaviour, fate and health risks of contaminants
918 in the end-products of freeze-drying and alkaline hydrolysis are still limited. Hence, comparative
919 research is needed to investigate these aspects relative to conventional funeral technologies.

920

921 **5.2.3 Engineering controls**

922 This encompasses measures aimed at reducing human exposure to risks or the removal of
923 the hazards through the use of technology or engineering interventions. This may include
924 automating highly hazardous procedures to reduce human exposure. Other engineering controls
925 include application of technologies for the treatment of solid wastes, wastewaters, contaminated
926 drinking water and air-borne particulates. The removal of organic contaminants in wastewater and
927 contaminated drinking water sources can be achieved using several treatment processes. These
928 include; adsorption, chlorination, advanced oxidation processes, boiling, solar disinfection
929 (SODIS), membrane filtration and ceramic filters (Pooi and Ng, 2018, Table 3).

930 Table 3 presents as summary of the various treatment processes, including their advantages
931 and limitations. The methods for removal of organic contaminants have been the subject of several
932 papers, including reviews (Pooi and Ng, 2018, Michael et al., 2020, Table 3). These treatment
933 processes provide critical tools for reducing contaminant loads in water and wastewater. The
934 choice of method will depend on several factors, including water/wastewater treatment objectives,

935 installation and operation costs, and availability of technology and expertise. Compared to other
936 treatment processes, the capacity of low-cost drinking water treatment methods such as boiling
937 and solar disinfection to remove TOCs is the least studied. Hence, it remains unclear which low-
938 cost methods are most effective in removing organic contaminants in drinking water.

939

940

941

942 **Table 3.** Summary of treatment processes for the removal of organic contaminants in aqueous systems.
 943

Process	Summary of process and principles	Benefits and advantages	Limitations and remarks	Review
Adsorption	Removes a wide range of organic contaminant via physico-chemical sorption	Low-cost method. Several natural and synthetic adsorbents exist	Only transfers contaminants from one phase (aqueous) to solid phase and generates large quantities of sludge	Grassi et al., 2012;
Chlorination	Removes micro-organisms including antimicrobial resistant bacteria and resistance genes via. Limited capacity to remove chemical organic contaminants.	Widely used method for disinfection in both centralized and decentralized systems	May react with organic matter to generate carcinogenic disinfection by-products	Tak and Kumar, 2017
Advanced oxidation processes	Uses strong oxidants (O ₃ , H ₂ O ₂) to generate reactive radicals that degrade organic contaminants. Often combined with ultraviolet irradiation (UV) and photocatalysts (e.g., TiO ₂) to enhance removal efficiency.	Highly effective in degrading organic contaminants. Organic contaminants can be completely mineralized to relatively benign by-products.	Have high energy and chemical requirements, hence relatively expensive compared to other methods. May not be ideal in low-income countries due to high costs.	Michael et al., 2020; Zhou et al., 2020
Membrane filtration	Relies of size exclusion and contaminant size to remove from from aqueous systems. Processes include reverse osmosis, nano-filtration among others	Very high removal efficiencies for various contaminants. Have limited effects of organoleptic properties of water.	Prone to several fouling processes such as scaling, and relatively expensive. Limited use of the technology in developing countries due to high costs.	Pendergast and Hoek, 2011; Chang et al., 2019
Boiling	Disinfects water by using high temperatures to destroy cell components. Volatile organic contaminants may be removed via volatilization.	Low-cost disinfection methods commonly used in developing countries and humanitarian emergencies for drinking water treatment	May increase contaminant concentrations via evaporation. Limited data exist on removal of antimicrobial resistance and emerging contaminants.	Gwenzi et al., 2018
Solar disinfection (SODIS)	Relies on UV irradiation and high temperature to destroy microbial cell components and organic contaminants via photodegradation. Similar to UV violet irradiation.	Low-cost method used for drinking water treatment in developing countries. Mainly targeted for removal of pathogens.	Requires long exposure times. Limited capacity to remove organic contaminants. Limited research on removal of antimicrobial resistance and emerging contaminants.	Pooi and Ng, 2018; Pichel et al., 2019
Biosand filtration	Formation of a biofilm or <i>Schmutzdecke</i> critical for the removal of pathogens via slow filtration, and subsequent predation and/or die-off.	Low-cost method commonly used for drinking water treatment in developing countries and humanitarian emergencies.	Low capacity to remove dissolved contaminants, and may promote antimicrobial resistance via biofilms. Limited data on removal of emerging contaminants.	Pooi and Ng, 2018
Ceramic filters	Relies on size exclusion and antimicrobials (e.g., Ag nanoparticles) to remove pathogens.	Low-cost methods for treatment of drinking water in developing countries especially pathogens.	Treat small volumes, and Ag may occur in treated water. Limited data on removal of emerging contaminants.	Pooi and Ng, 2018

945 **5.2.4 Regulatory and policy perspectives**

946

947 Despite the occurrence of TOCs and the putative health risks, consensus exists that,
948 globally, the funeral industry is under-regulated (Horton, 2003; Davidson and Benjamin, 2006;
949 Miranda, 2016; Kuchniski et al., 2019). Specifically, evidence from the USA, Canada, Europe and
950 Africa shows that the regulation and control of the funeral industry do not commensurate with the
951 health risks (Horton, 2003; Davidson and Benjamin, 2006; Chiapelli and Chiapelli, 2008; Miranda,
952 2016; Kleywegt et al., 2019; Kuchniski et al., 2019; Turajo et al., 2019; Zume et al., 2019; Gwenzi,
953 2020). For example, compared to other hazardous wastes from the health care industry, a cursory
954 regulatory and research attention has been paid to the disposal of human cadavers, solid waste and
955 wastewaters from the funeral industry (Gwenzi, 2020). Thus, in several countries, thanatopraxy
956 care facilities, cemeteries and crematoria are often excluded in routine environmental surveillance
957 systems. This stems from the fact that, in most countries (e.g., USA, Zimbabwe), the funeral
958 industry falls under commercial service providers, which is less regulated compared to the health
959 care industry (Davidson and Benjamin, 2006; Gwenzi, 2020).

960 Regulatory inconsistencies also exist; for example, on the one hand, blood, body fluids and
961 even solid wastes from the health care systems are classified as hazardous wastes, and are handled
962 and disposed accordingly. Yet on the other hand, the same blood and body fluids in thanatopraxy
963 care are directly discharged into septic systems or the municipal wastewater systems without prior
964 treatment (Chiapelli and Chiapelli, 2008). In most countries, embalming now appears to be the
965 norm rather than the exception. Yet evidence shows that, barring a few exceptions, embalming is
966 not a legal requirement (e.g., USA, Chiapelli and Chiapelli, 2008; Europe, Miranda, 2016). The
967 exceptions are long-distance transport of human cadavers and those used for medical scholarship
968 (Chiapelli and Chiapelli, 2008; Varlet et al., 2019). Thus, the indiscriminate embalming appears to
969 be driven by financial benefits derived from such a practice by the funeral industry, rather than
970 being a scientific or legal justification. This further highlights the weak regulatory framework.
971 Collectively, these scenarios point to the need to develop robust and appropriate regulatory and
972 policy frameworks for the funeral industry.

973 The regulatory and policy frameworks may entail developing a dual classification system
974 for the funeral industry: (1) the business aspects including funeral assurance can be classified under
975 commercial service providers just like medical insurance, while, (2) the disposal of human
976 cadavers, embalming products, wastes and wastewaters from the funeral industry should be
977 governed by regulations similar to those of the health care industry. Therefore, the regulatory and
978 policy frameworks should be premised on the fundamental fact that, human cadavers, embalming
979 products, solid wastes, wastewaters and air-borne particulates from the funeral industry are
980 'hazardous'. This notion is similar to the one used for medical wastes from the health care industry.
981 Yet there is need for caution to avoid classifying human cadavers as 'hazardous wastes', because
982 this will be viewed as insensitive and inhumane in some socio-cultural settings.

983 Embracing the fact that human cadavers, embalming chemicals and associated wastes and
984 wastewaters are 'hazardous' has several implications. First, the production, marketing, purchase,
985 storage, use and final disposal embalming products will require a special permit. Such special
986 permits may specify the need to maintain records on existing stocks, number of people embalmed,
987 and disposal of wastes and wastewaters. In addition, such permits may also specify the following
988 requirements: (1) mandatory accreditation and registration, including the minimum qualifications
989 of embalmers and funeral directors, (2) best safety, health and environment procedures, including
990 the use of appropriate labelling and insignia to communicate hazards, and (3) proper management

991 practices for solid wastes, wastewaters and air-borne emissions (LaGrega et al., 2001). This will
992 ensure that, regulatory agencies regard the funeral industry among other high-risk industries (e.g.,
993 health care system) that generate hazardous wastes. In turn, funeral industry will be accorded the
994 appropriate routine environmental, occupational and public health surveillance and control.

995 The notion of ‘hazardous’ also calls for proper planning, design, operation and monitoring
996 of engineered systems for the disposal of human cadavers, wastes and wastewaters. In this regard,
997 a ‘new generation’ of properly engineered cemeteries will be required. The design principles of
998 such cemeteries could be adapted from that of engineered sanitary landfills (Qian et al., 2001).
999 This include the need for; (1) runoff and drainage collection system, (2) hydraulic liners or barriers
1000 to restrict leachate migration into groundwater systems, and (3) leachate monitoring and collection
1001 system, and (4) on-site wastewater treatment system for treating contaminated leachate, drainage
1002 and runoff before discharge into the environment or sewer systems. Moreover, mandatory
1003 incineration of solid waste and on-site wastewater treatment will be required. Incinerators and
1004 crematoria should be equipped with emission control and monitoring systems.

1005 The formulation of regulatory and policy frameworks requires a strong scientific evidence
1006 base, which is currently missing in developing regions, including Africa. Moreover, environmental
1007 and public health surveillance systems are either weak or non-existent (Loewenson, 1995). As with
1008 other health risks such as infectious diseases, local capacity and expertise to develop effective
1009 regulatory and policy frameworks, and surveillance and control systems in the funeral industry are
1010 currently lacking in developing countries. Thus, the international community, through partnerships
1011 with relevant international agencies (e.g., WHO, UNESCO, UNICEF) and developed countries
1012 could assist with the following: (1) the development of effective surveillance and control systems,
1013 and (2) facilitate and provide guidance in the formulation of regulatory and policy frameworks.
1014 This can be achieved through the following: (1) the provision of experts/specialists, (2) capacity-
1015 building initiatives, including training workshops, and (3) provision and mobilization of resources
1016 and equipment for conducting research to develop the local scientific evidence base.

1017

1018 **6 Future research and perspectives**

1019 Future research should address key knowledge gaps in several thematic topics on TOCs
1020 in the funeral industry. The application of conventional and emerging research tools is critical in
1021 addressing these knowledge gaps (*Section 6.2*).

1022

1023 **6.1 Future Research**

1024 (1) *Occurrence and partitioning of TOCs*

1025 Limited data exist on the occurrence of TOCs in various reservoirs such as solid wastes,
1026 wastewaters and air-borne particulates in the funeral industry. Thus, further studies are needed to
1027 determine the occurrence and partitioning of TOCs and their metabolites in the various reservoirs.
1028 Such studies should estimate inventories of annual emission of TOCs in solid wastes, wastewater
1029 and air-borne particulates in various hotspot sources in the funeral industry.

1030 (2) *Behaviour and fate of TOCs*

1031 Limited data exists on the behaviour, fate and degradation kinetics of TOCs in tropical
1032 environments predominant in developing countries. Thus, comparative studies are required to
1033 understand the biogeochemical behaviour, degradation kinetics and fate of TOCs in tropical versus
1034 temperate environments. Speciation modelling may provide insights on the degradation kinetics,
1035 while chronosequence studies on cemeteries may provide cues on the long-term fate.

1036 (3) *Ecotoxicology of TOCs*

1037 Limited data exists on ecotoxicology of organic contaminants at environmentally relevant
1038 concentrations such as in solid wastes, wastewaters and air-borne particulates in funeral industry.
1039 In addition, the health risks arising from mixtures of TOCs and their interaction with other health
1040 stressors are poorly understood. Studies based on toxicokinetic, pharmacokinetic and energy
1041 budgets modelling (Dietz et al., 2015; Sonne et al., 2014, 2015) are needed to understand the
1042 uptake, bioaccumulation, (bio)transformation and fate of TOCs in biota.

1043 (4) *Ecological health risks of organic contaminants*

1044 Evidence on the ecological health effects of emerging organic contaminants remain largely
1045 inferential, while direct evidence remain weak. Thus, studies are required to better understand the
1046 ecological health risks at various level of biological organization using a combination of
1047 conventional and emerging tools (**Section. 6.2**).

1048 (5) *Removal of TOCs by low-cost water treatment processes*

1049 Literature on the removal of TOCs is limited to conventional and advanced treatment
1050 methods (Michael et al., 2020). Data on the removal of TOCs by low-cost water treatment methods
1051 such as slow biosand filtration, boiling and solar disinfection are scarce. Such low-cost methods
1052 are routinely used in humanitarian emergencies and by low-income communities. Such
1053 information is critical for the choice of methods for drinking water treatment in such settings.

1054 (6) *Quantitative human health risk assessment*

1055 Evidence on the human health risks of TOCs in the funeral industry and other
1056 environmental reservoirs is predominantly inferential. The dominant pathways contributing to
1057 human exposure and health risks in the funeral industry remain unknown. Hence, systematic case-
1058 control studies in human toxicology and epidemiology are required to directly establish the link
1059 between the occurrence of TOCs and incidences of specific human health outcomes.

1060

1061 **6.2 Moving forward in health risk assessment**

1062 To this point, it is evident that, receptor organisms including humans are co-exposed to a
1063 milieu of contaminants and health stressors. Yet most environmental risk assessments, including
1064 those conducted for as part of the pre-registration and approval of synthetic chemicals rarely
1065 consider mixtures and the their interactions with other health stressors. Moreover, understanding
1066 the ecological and human health risks of TOCs in environmental systems is quite complex.

1067 The limitations inherent in the current evidence and risk assessment protocols calls for
1068 caution when drawing conclusion on health risks based on single chemical bioassays. Specifically,
1069 this raises the question, ‘*Do current environmental guideline limits based on existing health risk
1070 assessments really protect ecological and human health, and if not, how do we move forward from
1071 here?*’ This question resonates with earlier concerns raised in the USA (Hayes et al., 2016) and
1072 the European Community (Carvalho et al., 2014), questioning whether or not current
1073 environmental regulations and their guideline limits were sufficient to safeguard human and
1074 ecological health. This call for the urgent attention of the research community, regulators, funding
1075 agencies and policy makers to reflect on current evidence and risk assessment frameworks.

1076 The development and subsequent application of the next generation tools for understanding
1077 the health risks of mixtures and their interactions with other health stressors constitute the new
1078 research frontier in (eco)toxicology and health risk assessment. Such future studies should address
1079 health risks at various level of biological organization, including molecular, organ, individual,
1080 population, community, ecosystem scales and even ecosystems goods and services. Conventional
1081 research tools for (eco)toxicology, health risk assessment and epidemiology include; (1)
1082 manipulative experiments using microcosms, mesocosms and even real ecosystems (van Donk et

1083 al., 2016), coupled with hierarchical modelling (Chen et al., 2013), (2) protocols for
1084 ecotoxicological and health risk assessments (e.g., EC, 2006; OECD, 2015; US EPA, 2020a, b),
1085 and (3) tools in human toxicology and epidemiology of contaminants, including case-control
1086 experiments (He and Huang, 2010; Drakvik et al., 2020). However, in view of the limitations
1087 highlighted, these tools only provide a starting point, but there is need to improve and adapt them.

1088 Recent advances in analytical techniques and data analytics provide emerging tools and
1089 unprecedented opportunities to better understand the health risks of TOCs (Table 4). These tools
1090 include; (1) advanced and highly sensitive analytical techniques including those for 2-D and 3-D
1091 imaging (Lindner et al., 2015; Mehrian et al., 2020), (2) isotopic labelling (Boecklen et al., 2011;
1092 Parnell et al., 2013), (3) molecular/genomic tools (Pulojar et al., 2012; Valli et al., 2020), (4) big
1093 data analytics such as machine learning and artificial intelligence (Vestergaard et al., 2017; Hyun
1094 et al., 2020), and (5) *in silico* or computational techniques (Vuorinen et al., 2013; Raies and Bajic,
1095 2016; Nwaimu and Aduba, 2020). Surprisingly, despite the immense opportunities offered by
1096 emerging tools, their application has received limited attention among the research community
1097 focusing on ecotoxicology and health risks of contaminants. Table 4 presents a summary of the
1098 various emerging tools and their potential applications in understanding the behaviour,
1099 (eco)toxicology and health risks of TOCs.

1100 It is envisaged that, future research integrating conventional and emerging tools will lead
1101 to the development more robust ‘next generation’ health risk assessment protocols, and
1102 surveillance and control systems. Given the multi-disciplinary nature of health risks of TOCs, such
1103 future research should bring together diverse disciplines, including; environmental scientists,
1104 toxicologists, ecologists, epidemiologists, and experts in mathematics and computer sciences.

1105
1106

1107 **Table 4.** Potential applications of merging tools to understand the behaviour, (eco)toxicology and health risks of toxic organic
 1108 contaminants (TOCs).

Emerging tools	Examples	Potential applications and remarks	References
(1) Advanced and sensitive analytical methods	(a) HPLC-ICP-MS, Laser ablation ICP-MS, (b) Computed X-ray tomography	(a) Detailed characterization of the concentrations and speciation of TOCs and their metabolites. (b) Non-invasive 2-D and 3-D imaging to understanding internal structure of complex TOC-solid matrix interactions in soils and sediments.	Lindner et al., 2015 Mehrian et al., 2020.
(2) Isotopic labelling	Use of TOCs labelled with ¹⁵ N and ¹³ C isotopes	Understanding the environmental behaviour, partitioning and fate of TOCs, including elimination mechanisms, uptake, bioaccumulation, biomagnification, (bio)transformation and trophic transfers.	Boecklen et al., 2011; Parnell et al., 2013
(3) Genomics/molecular techniques	(a) DNA probing, RNA sequencing, polymerase chain reaction (PCR) (b) metabolomics, proteomics, transcriptomics, metagenomics	(a) Understanding effects of TOCs on specific gene expressions via gene profiling, including those related to oxidative stress. (b) Can be coupled to network analysis to disentangle complex ecological processes, including trophic interactions, metabolic networks and ecosystem functions.	Pulojar et al., 2012 Valli et al., 2020
(4) Big data analytics	Machine learning (ML), artificial intelligence (AI), data mining (DM)	(a) Big data analytics have unique capacity for analysis, integration, synthesis and visualization of data from multiple sources (i.e., big data), including surveillance data on TOCs in environmental media, health stressors, (eco)toxicology, health risks, human toxicology and epidemiology, and their interactions (b) Machine learning, artificial intelligence and data mining tools are ideal for the prediction and extraction of spatio-temporal patterns and trends from 'big data'	Vestergaard et al., 2017; Hyun et al., 2020).
(5) <i>In silico</i> or computational techniques	Toxicokinetic modelling, network analysis, scenario modelling	(a) Rapid design and analysis of experiments to gain early insights that guide laboratory and field experimentation. (b) Complement isotopic labelling studies to understand speciation, behaviour and fate of TOCs, including bioaccumulation and (bio)transformation. (c) Ideal for scenario analysis modelling to investigate 'what if' questions in (eco)toxicology and health risk assessments.	Vuorinen et al., 2013; Raies and Bajic, 2016)

1109

1110 In this research effort, international, regional and national environmental agencies (UNEP, OECD,
1111 US EPA) with experience in health risk assessment should provide the lead, in collaboration with
1112 their developing world counterparts. For now, safeguarding ecological and human health in the
1113 face of imperfect knowledge may require adopting the precautionary principle.

1114

1115 **7 Conclusions and Outlook**

1116

1117 The current review investigated the nature, occurrence and health risks of TOCs in the
1118 funeral industry. Data show that the funeral industry is a continuum of several hotspots of TOCs,
1119 including autopsy, thanatopraxy care facilities, cemeteries and crematoria. TOCs detected include;
1120 embalming products, persistent organic pollutants, synthetic pesticides, pharmaceuticals, personal
1121 care products and illicit drugs. Hotspot reservoirs of TOCs in the funeral industry include; human
1122 cadavers, solid wastes, wastewaters and air-borne particulates and aerosols. Surface and sub-
1123 surface hydrological processes control the transfer and dissemination of TOCs in environmental
1124 systems. Occupational exposure may occur in autopsy, thanatopraxy, cemeteries and crematoria
1125 via ingestion, inhalation of air-borne particulates and aerosols, and dermal intake via cuts and
1126 wounds. Non-occupational human exposure occurs via ingestion of contaminated drinking water,
1127 and aquatic and marine foods. Inhalation of volatile TOCs and dermal contact during body
1128 viewing, washing and handling of human cadavers during home care, funerals and burials may
1129 also occur.

1130 Health risks of TOCs are quite diverse, and dependent on the nature of contaminants. The
1131 health risks include; (1) acute and chronic toxicity such a genotoxicity, teratogenicity, endocrine
1132 disruption and neurodevelopmental disorders, (2) antimicrobial resistance induced by
1133 pharmaceuticals and co-selecting emerging contaminants, (3) info-disruption caused by emerging
1134 contaminants via blocking and biomimickry of natural info-chemicals, and (4) disruption of
1135 ecosystem functions, including trophic interactions and biogeochemical cycling. However, barring
1136 formaldehyde, and inferential evidence, systematic quantitative studies directly linking TOCs in
1137 the funeral industry to specific health outcomes are still limited. Hence, the limitations of current
1138 evidence, and health risk assessment protocols were discussed. These limitations point to the need
1139 to develop and apply the next generation of better health risk assessment tools. To safeguard
1140 ecological and human health, a framework for hazard identification, risk assessment and mitigation
1141 (HIRAM), including prevention and control was developed. Future research directions, including
1142 key knowledge gaps, and opportunities presented by several emerging tools were highlighted.

1143

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1148

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1152

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